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### **The extensive mercury contamination in soil and legacy sediments of the Paglia River basin (Tuscany, Italy): interplay between Hg-mining**

Questa è la versione Preprint (Submitted version) della seguente pubblicazione:

*Original Citation:*

The extensive mercury contamination in soil and legacy sediments of the Paglia River basin (Tuscany, Italy): interplay between Hg-mining waste discharge along rivers, 1960s economic boom, and ongoing climate change / Fornasaro S.; Morelli G.; Rimondi V.; Fagotti C.; Friani R.; Lattanzi P.; Costagliola P.. - In: JOURNAL OF SOILS AND SEDIMENTS. - ISSN 1614-7480. - STAMPA. - 22:(2022), pp. 656-671. [10.1007/s11368-021-03129-0]

*Availability:*

This version is available at: 2158/1290866 since: 2022-11-21T09:17:04Z

*Published version:*

DOI: 10.1007/s11368-021-03129-0

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(Article begins on next page)

# Journal of Soils and Sediments

## The extensive mercury contamination in soil and legacy sediments of the Paglia River basin (Tuscany, Italy): interplay between Hg-mining waste discharge along rivers, 1960s economic boom, and ongoing climate change

--Manuscript Draft--

<b>Manuscript Number:</b>	JSSS-D-21-00811R1
<b>Full Title:</b>	The extensive mercury contamination in soil and legacy sediments of the Paglia River basin (Tuscany, Italy): interplay between Hg-mining waste discharge along rivers, 1960s economic boom, and ongoing climate change
<b>Article Type:</b>	Research Article
<b>Section/Category:</b>	Sediments
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<b>Order of Authors Secondary Information:</b>	
<b>Funding Information:</b>	
<b>Abstract:</b>	<p><b>Purpose</b> The extensive Hg contamination in soil and sediments occurring along the Paglia River (Central Italy) is the result of the interplay between the geomorphological changes of the river and the anthropic activities, primarily associated to the exploitation of Hg-deposits in the Monte Amiata mining district (MAMD). The present study points out the implications of the morphological changes occurred in the last 200 years of the Paglia River on the distribution of Hg along the floodplain and riverbed, which today represent one of the main Hg-reservoirs in the MAMD.</p> <p><b>Materials and Methods</b> The temporal changes of the Paglia riverbed and the extent of its alluvial deposits were reconstructed by a GIS-based analysis of the available maps and aerial photos. The Hg-concentration in soil and sediment samples, collected along five transects transverse to the Paglia River course, was determined by ICP-MS.</p> <p><b>Results and Discussion</b> Samples along the investigated Paglia River segment typically show Hg-contents exceeding the Italian threshold for residential and public green soil use (1 mg kg<sup>-1</sup>). The distribution of Hg in the Paglia floodplain results from the combination of exceedance of sediment yield to the river during mining activities, that fed the floodplain with large amounts of Hg-contaminated sediments during its braided stage</p>

	<p>about 100 years ago, and the morphological changes of the river, that led to the evolution from a braided to the present-day single channel river.</p> <p>The magnitude of the extension of Hg-contamination, the river geomorphologic changes, and the processes of transport, deposition, and re-suspension did not allow a natural “clean up” of the river system, which shows a low resilience. Under high flow conditions, and especially in coincidence with intense rain events, large amounts of Hg stored in the overbank sediments are mobilized and redistributed, contributing to make the floodplain a secondary Hg-source. Extreme weather events, expected to intensify as a consequence of climate change, will contribute to the recurrent distribution of Hg-contaminated legacy sediments in the floodplain and along the Paglia river course.</p> <p>Conclusion</p> <p>From a water/land management perspective, the variability of the river flow, associated with an increase of extreme flood events driven by climate change, will affect the distribution of Hg-contaminated particles in the Paglia River, contributing to the Hg input into the Mediterranean Sea in the future.</p>
<p><b>Response to Reviewers:</b></p>	<p>Comments from the Editor</p> <p>Paglia River always with capital letter R</p> <p>The text was modified according to the suggestion of the Editor.</p> <p>Paglia River and Tibera River should be preferred to Paglia R. and Tibera R.</p> <p>The text was modified according to the suggestion of the Editor.</p> <p>"overbanks" should be written as "overbank sediments"</p> <p>The term was modified where deemed necessary.</p> <p>units should be written with powers in superscript, e.g., mg kg<sup>-1</sup> (with <sup>-1</sup> as superscript), not mg/kg</p> <p>The text and the figures were modified according to the suggestion of the Editor.</p> <p>“braidplain” should be considered instead of floodplain, where the term is more appropriate e.g., in Fig. 2</p> <p>We believe it is more appropriate to keep the generic term “floodplain”, because in the study area the suggested “braidplain” would be applicable only in the pre-mining period.</p> <p>Reviewer #1</p> <p>While I do not have a background in river morphology, I found the paper compelling and well written. I support its publication. My one critique it that while the title has "climate change" in it, and the body of the paper and the conclusion discuss climate change, the theme does not show up in the abstract section. This problem needs to be addressed.</p> <p>As suggested by the reviewer, we emphasized in the abstract the hypothesized relation to climate change (lines 33-35). Moreover, we added a statement at the end of the Introduction (lines 83-84)</p> <p>Reviewer #2</p> <p>This paper entitled “The extensive mercury contamination in soil and legacy sediments of the Paglia River basin (Tuscany, Italy): interplay between Hg-mining waste discharge along rivers, 1960s economic boom, and ongoing climate change” represents interesting study on temporal changes in contamination of floodplain soils and sediments in former mining watershed. It brings new insight onto the interplay between mine production/economical situation, river stages and climatic variables. But major revisions of certain sections are needed so that this paper has context for an international reader and complies with international standards.</p> <p>Apart of the minor issues listed below, I have serious problem with the section 2.4 Geochemical analysis. This section should be expanded and described in much greater detail. What exactly means that "concentrations were recalculated to the grain size fraction &gt;2 mm" and why was this done? This should be explained for reader.</p> <p>We are aware that this point of recalculation is a (probably useless) complication. However, it is a specific requirement of the Italian legislation (DL 152, 2006 “Norme in materia ambientale”), and as such cannot be ignored by environmental agencies. As stated in the Acknowledgments, the study arises from an agreement between the University and the regional environmental agency (ARPAT) and it was ultimately funded by the regional government. In practice, according to the Italian legislation</p>

requirement, the concentration of a given pollutant, actually measured in the fraction <2 mm, has to be normalized to the total weight of the sample (including the fraction >2mm). This requirement is based on the assumption that the fraction > 2mm does not contain significant amounts of the pollutant, an assumption that may be unjustified in some cases, but, as demonstrated in a previous study (Colica et al., 2019), is essentially correct for Hg in the Paglia River sediments. Moreover, in samples collected for this study the fraction >2mm is virtually absent. In conclusion, in this specific case the observance of law requirements does not change the results. We added a phrase in the text stating that the >2mm correction is negligible (lines 192-194)

Furthermore, leaching experiments - why did you do those? I could not find result/comment or discussion on these analyses. Where are these results evaluated? The only place where they can be found is Table 2. I also have some doubts on leaching experiments e.g., what kind of Hg do you think has been mobilized from the sediment by distilled water? Was the eluate filtered? Were selected samples replicated to have an idea about the replicability of leaching? Why are the results of leaching in ug/L (Tab. 2)? This should be recalculated onto solid phase... so that it is directly comparable to the total Hg.

Once and again, the methodology described for the leaching experiments is a law requirement. Indeed, previous studies (e.g., Rimondi et al., Chem. Geol. 380, 2014) suggest that most Hg in this kind of samples occur as extremely low solubility phases (sulfides), so leaching tests with pure water can only remove minimal amounts of (sparingly) soluble phases (e.g., Hg<sup>0</sup>). The ug/l unit is again according to the law; it can be easily recalculated onto the solid mass. But, overall, we agree with you that these data are of little significance for the paper. Therefore, we removed any mention of these data

Same issue, for the total analysis in the digests - was the repeatability tested? Accuracy was tested by independent standards what are these? Do you mean reference materials (CRM)? If you do, specify which exactly did you use, how many replicates did you make and how precise was the analysis.

At last, you indicate 10% analytical precision - this is rather high. EPA standards for Hg analysis usually aim to 5%? Please comment on this.

Thank you for raising this issue and helping us to clarify this point. We expanded the related paragraphs (lines 192 and following), hopefully answering to your questions. Concerning analytical precision, we point out (as stated in lines 201-203) that we refer to the difference between analyses of different aliquots of the same bulk sample. Precision is usually better than 10%, but in few cases (probably because of the well-known "nugget effect") is up to 10%. In any case, we believe it does not affect the overall significance of the obtained data.

Please consider moving Table 2 into supplement after editing it with respect to my previous remark. The new Hg data are included in figs 4 and 5. Furthermore Table 2 is only referenced once through the whole paper, but it is quite extensive. We accepted the suggestion

Figure captions in the Figures section of PDF appear on different pages with different figures... e.g., page 27 and 29, make sure these appear where they should as it complicates work with PDF.

Thanks for the comment. During the resubmission process, we will pay attention to the correct layout of the PDF.

Line 96 m asl - please unify this abbreviation so it is consistent through the whole paper as "m a.s.l."

The text was modified according to the suggestion

Line 119 what exactly does variable flow mean? Please explain to reader. We added new information in order to clarify the concept.

Line 256 Thanks for the paragraph above. Very interesting. I wish you indicate at the end when exactly was the complete halt - which year or years. We added specific details in the text (line 270)

Line 271 channel reduction... was this natural or anthropogenic process?

As explained in the following lines 287-288, the channel reduction was mainly due to anthropogenic processes (e.g., recovery of land for agriculture, river management works, land-use changes, building of weirs and dams, quarrying of sediments).

Line 280 which Italian rivers are we talking about here, be particular  
We added new information in the text (lines 294-295)

Line 282 through the whole paper "World War I" and "World War II" are referred to this way.

The text was modified according to the suggestion

Line 287-288 I do not understand this sentence. How does can be "contaminated sedimentary bodies suspended"? Please rewrite.

We rewrote the sentence.

Line 293-294 I do not understand this sentence, especially the end of it about floodplain of 1954.

We rewrote the sentence.

Line 299-300 Great! We have seen a lot of that here.

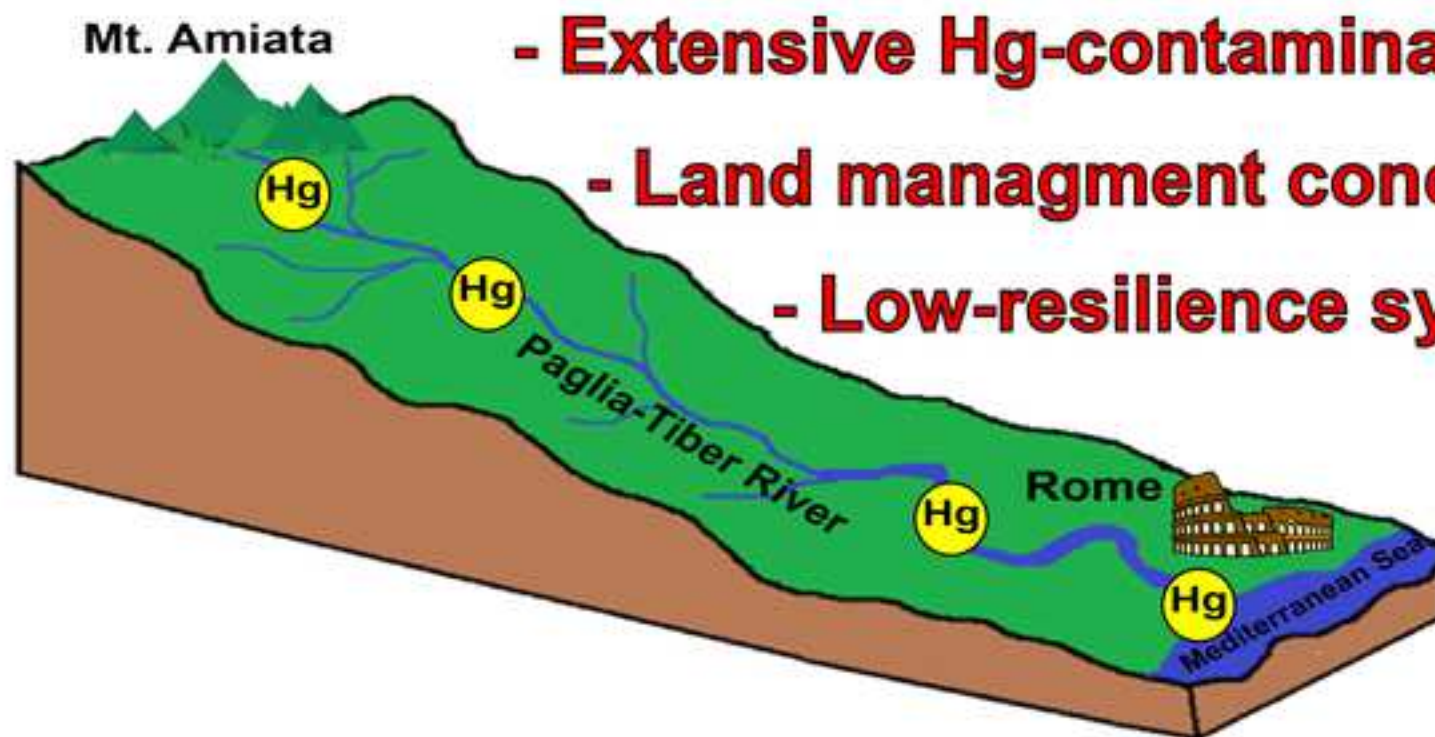
Thank you for the comment

Line 375-376 contraction of snow periods of snow affect river dynamics but how? The reader should be informed.

The sentence was removed.

Line 382 Good. I acknowledge these statements on conservation measures.

Thank you for the comment.



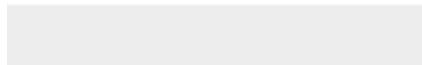
- Extensive Hg-contamination
- Land management concerns
- Low-resilience system



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**Supplementary Material**

Supplementary\_Information\_revised.docx



1 **The extensive mercury contamination in soil and legacy sediments of the Paglia River basin (Tuscany,**  
2 **Italy): interplay between Hg-mining waste discharge along rivers, 1960s economic boom, and ongoing**  
3 **climate change**

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9

10 **Abstract**

11 **Purpose**

12 The extensive Hg contamination in soil and sediments occurring along the Paglia River (Central Italy) is the  
13 result of the interplay between the geomorphological changes of the river and the anthropic activities, primarily  
14 associated to the exploitation of Hg-deposits in the Monte Amiata mining district (MAMD). The present study  
15 points out the implications of the morphological changes occurred in the last 200 years of the Paglia River on  
16 the distribution of Hg along the floodplain and riverbed, which today represent one of the main Hg-reservoirs  
17 in the MAMD.

18 **Materials and Methods**

19 The temporal changes of the Paglia riverbed and the extent of its alluvial deposits were reconstructed by a  
20 GIS-based analysis of the available maps and aerial photos. The Hg-concentration in soil and sediment  
21 samples, collected along five transects transverse to the Paglia River course, was determined by ICP-MS.

22 **Results and Discussion**

23 Samples along the investigated Paglia River segment typically show Hg-contents exceeding the Italian  
24 threshold for residential and public green soil use (1 mg kg<sup>-1</sup>).

25 The distribution of Hg in the Paglia floodplain results from the combination of exceedance of sediment yield to  
26 the river during mining activities, that fed the floodplain with large amounts of Hg-contaminated sediments  
27 during its braided stage about 100 years ago, and the morphological changes of the river, that led to the  
28 evolution from a braided to the present-day single channel river.

29 The magnitude of the extension of Hg-contamination, the river geomorphologic changes, and the processes  
30 of transport, deposition, and re-suspension did not allow a natural “clean up” of the river system, which shows  
31 a low resilience. Under high flow conditions, and especially in coincidence with intense rain events, large  
32 amounts of Hg stored in the overbank sediments are mobilized and redistributed, contributing to make the  
33 floodplain a secondary Hg-source. Extreme weather events, expected to intensify as a consequence of climate  
34 change, will contribute to the recurrent distribution of Hg-contaminated legacy sediments in the floodplain and  
35 along the Paglia river course.

36 **Conclusion**

37 From a water/land management perspective, the variability of the river flow, associated with an increase of  
38 extreme flood events driven by climate change, will affect the distribution of Hg-contaminated particles in the  
39 Paglia River, contributing to the Hg input into the Mediterranean Sea in the future.



40

41 **Keywords**

42 legacy sediments; fluvial dynamics; mercury; Monte Amiata

43

44 **1. Introduction**

45 Geomorphic features of riverine systems result from the balance of many parameters (e.g., water and total  
46 sediment load; Schumm and Harvey 1999; Calle et al. 2017), that may in turn be affected by factors such as  
47 climate changes and human activities (Grabowski and Gurnell 2016; Marchamalo et al. 2016; Calle et al. 2017;  
48 Owens 2020; Vauclin et al. 2020). Specifically, in fluvial systems draining mining areas, mining activities may  
49 contribute significantly to the modification of river morphology, influencing sediment supply and the associated  
50 processes of erosion, transport, and deposition (e.g., Ciszewski and Grygar 2016; Davis et al. 2018). In  
51 addition, pollutants associated with mining particulate, such as heavy metals, are responsible of large-scale  
52 contamination up to several hundred kilometers away from the mining area (Martin and Maybeck 1979; Schafer  
53 et al. 2006; Mayes et al. 2013; Rimondi et al. 2019). Mining-contaminated legacy sediments deposited along  
54 waterways may remain stored within river channels and on floodplains for hundreds or thousands of years  
55 (Salomons and Förstner 1984; Macklin and Lewin 1989; Pavlowsky et al. 2017; Davis et al. 2018; Rimondi et  
56 al. 2019). They become diffuse sources of contamination if re-mobilized, for example by overbank erosion  
57 during flood events, or by human activities (e.g., gravel mining; Macklin et al. 1997; Pavlowsky et al. 2017;  
58 Colica et al. 2019). Floodplains therefore play an important role as both sinks and sources of metal  
59 contaminants in mined watersheds (Bradley 1989; Horowitz 1991; Lecce and Pavlowsky 1997; Coulthard and  
60 Macklin 2003; Ciszewski and Grygar 2016; Pavlowsky et al. 2017).

61 In the last 200 years, Italian waterways experienced considerable changes that triggered deep modifications  
62 of their original morphology (Surian and Rinaldi 2003; Cencetti et al. 2017, and references therein), similarly  
63 to other European rivers (e.g., Garcia-Ruiz et al. 2011; Debolini et al. 2015; Pavanelli et al. 2019). Incision and  
64 narrowing of the active channel were the most frequently observed modifications (Cencetti et al. 2017). The  
65 Tiber River (central Italy) and its tributary Paglia River are no exception and were affected by similar processes.  
66 The Paglia River (49 km length) is one of the right-side tributaries of the Tiber River. Its morphological changes  
67 play a key role for the delivery of Hg to the Mediterranean Sea, since the river directly collects the runoff from  
68 one of the largest Hg ore districts in the world, the Monte Amiata Mining District (MAMD). Previous works  
69 described the morphological changes occurred in different sections of the Paglia River (e.g., Cencetti et al.  
70 2017), and the pervasive distribution of Hg in river sediments and soils of the Paglia River floodplain,  
71 highlighting how fluvial dynamics contribute to transport Hg contaminated sediments up to 200 km downstream  
72 the MAMD (Colica et al. 2019; Rimondi et al. 2019).

73 The present study points out the implications of the morphological changes in the last 200 years of the Paglia  
74 River (Tuscan stretch) on the buildup of the fluvial overbanks, which today represent one of the main Hg  
75 reservoirs in the MAMD district. Geochemical data obtained during a sampling campaign conducted in 2020  
76 by the Regional Environmental Protection Agency of the Tuscany (ARPAT) and by the environmental  
77 mineralogy group at Dipartimento di Scienze della Terra (DST), Università di Firenze, complement previous

78 studies by Colica et al. (2019) and Rimondi et al. (2019). These new results were integrated with previous data  
79 to assess the spatial and temporal variability of Hg contamination in the Paglia River floodplain.  
80 Specifically, the aims of this study are: i) to understand how the geomorphological (natural and anthropogenic)  
81 changes control the Hg distribution in the Paglia River floodplain; ii) to verify the implications of flood events  
82 on Hg distribution and resilience of the river system, taking into account the potential consequences of climate  
83 changes.

84

## 85 **2. Materials and methods**

### 86 *2.1 Study area*

#### 87 *2.1.1 Tiber-Paglia River system*

88 The Tiber River is the third longest river in Italy, flowing through the city of Rome and into the Mediterranean  
89 Sea (Cattuto et al. 1988; Ciccacci et al. 1988; Fredduzzi et al. 2007; Cencetti et al. 2017). The adjustments in  
90 its course in the last 250 ky were caused by an interplay between glacio-eustasy, sedimentary processes and  
91 regional uplifts (Marra et al. 2019). It was subjected to anthropic pressure probably before the establishment  
92 of the Roman Empire (Salomon et al. 2017). The most evident modifications took place during the 20th century  
93 and in the last decades following the Italian economic expansion, peaked between 1950 and 1970, with the  
94 construction of dams, sediment mining, and changes in the agriculture practices. The morphological changes  
95 occurred in the Tiber River are like those observed elsewhere in the Apennine area (Cencetti and Tacconi  
96 2005; Rinaldi and Simon 1998; Surian and Rinaldi 2003).

97 The Paglia River is one of the right-side tributaries of the Tiber River and arises from the junction of Pagliola  
98 and Cacarello creeks (388 m a.s.l., Fig. 1). The former drains the main mining and metallurgical center of MAMD  
99 near the Abbadia San Salvatore township. The Paglia River starts its course southeastward along a gentle  
100 slope, reaches the border between Tuscan and Latium regions (266 m a.s.l.) about 15 km from the starting  
101 point, and it enters into the Tiber River south of the city of Orvieto (Fig. 1C).

102 The geological-geomorphological structure of the upper basin of the Paglia River is linked to the formation of  
103 the Apennines during the Tertiary, and the subsequent post-collisional events (Marroni et al. 2015). The  
104 geology of the basin from the base to the top includes (Fig. 1C; Marroni et al. 2015): i) Tuscan and Ligurian  
105 Units (Paleozoic – Lower Miocene); ii) marine, transitional, and continental sedimentary successions (Lower  
106 Pliocene – Quaternary); iii) volcanic and volcano-sedimentary successions (Upper Pliocene – Upper  
107 Pleistocene); iv) continental deposits-debris and alluvial deposit (Quaternary). The quaternary continental  
108 deposits are characterized by i) holocenic fluvial deposits, present along the Paglia River valleys and its main  
109 tributaries, consisting mainly of sandy-silty beds and pebbles. The Paglia River cuts through its alluvial  
110 deposits, locally forming various orders of terraces; ii) Pleistocene deposits: fluvial-lacustrine deposits, mainly  
111 formed by conglomerates with sandy-silty beds levels. These deposits are arranged on large terraces located  
112 at higher elevation (from 5 to 20 m) compared to the current course of the Paglia River (e.g., Colica et al.  
113 2019). The presence of substrates characterized by erodible lithologies (Mio-Pliocene deposits) contributed to  
114 erosion processes with the formation of peculiar morphologies such as the "*biancane*" and the "*calanchi*"  
115 (Ciccacci et al. 2009). The shape of the Paglia River basin and the trend of the hydrographic network are

116 closely correlated with the structural characteristics of the Radicofani and Cetona grabens, set on normal fault  
117 systems with NNW-SSE trend, and trend-transforming systems with WSW-ENE trend (Sani et al. 2016).  
118 This area is characterized by a Mediterranean temperate climate, with hot and dry summer and cold and rainy  
119 winter. The average annual temperature is 10.5°C (period 1953 – 2000), and the average annual precipitation  
120 is 1480 mm (over the period 1925 – 2000; Ciccacci et al. 2009). About  $\frac{2}{3}$  of the total annual precipitation is  
121 concentrated in the autumn-winter season (Ciccacci et al. 1988). The Paglia River flow regime is controlled by  
122 seasonal variability, ranging from 0.3 m<sup>3</sup> s<sup>-1</sup> to 26 m<sup>3</sup> s<sup>-1</sup> monthly average (Fredduzzi 2005), with an annual  
123 average discharge of 2.45 m<sup>3</sup>·s<sup>-1</sup> (Cencetti et al. 2017). The hydrological periods are difficult to define (Moretti  
124 et al. 1988), due to the torrential regime of the initial part of the Paglia River, that may rapidly reach very high  
125 flow, collecting water contributes of a high number of tributaries. In general, the lowest and the highest water  
126 flow levels were recorded at the end of the summer period (September-October) and in winter-spring  
127 (November-March; Rimondi et al. 2014), respectively. An increase in mean monthly discharge was observed  
128 since 2003, due to a higher frequency of extreme flood events (Pattelli et al. 2014; Rimondi et al. 2014; Cencetti  
129 et al. 2017). This trend peaked with the flood of 2012 (mean monthly discharge: 91.7 m<sup>3</sup>·s<sup>-1</sup>, peak flow: 2663  
130 m<sup>3</sup> s<sup>-1</sup>; Cencetti et al. 2017).  
131 The Paglia River is physically shaped by sequential seasonal events of flooding and drying over a yearly cycle  
132 (Gasith and Resh 1999), reflecting the highly irregular rainfall patterns, with marked differences between wet  
133 and dry seasons, as most Mediterranean rivers. The concomitance of intense rainy days after dry summer  
134 periods, coupled with the scarce vegetation in the area, causes flash floods, with associated sliding-like mud  
135 and debris flows (Di Tria et al. 1999).

### 136 137 *2.1.2 The Monte Amiata mining district (MAMD)*

138 The MAMD district covers an area of ~400 km<sup>2</sup> and includes 42 former mines and 4 Hg roasting plants (Ferrara  
139 et al. 1998). The MAMD produced about 102,000 t of Hg between 1860s and 1980s (Colica et al. 2019),  
140 representing the third cumulative production ever reached in the world (Rimondi et al. 2015). The on-site  
141 metallurgical processing of cinnabar, the principal ore mineral, produced wastes, called calcines, with  
142 significant residual Hg contents (25 – 1500 mg kg<sup>-1</sup>; Rimondi et al. 2012; 2015). These wastes were often  
143 abandoned or discharged directly into the rivers adjacent to the mines, or used as filling material for road  
144 networks, house foundations, or landfills in crops (unpublished report, item T-1268 of the archives of the  
145 exploration company RIMIN). Numerous studies highlighted the environmental impact caused by over a  
146 century of mining and metallurgical activities in the MAMD, and the consequent contamination of Hg in the  
147 sediments transported by the Paglia River (Rimondi et al. 2019, and references therein). The first studies  
148 concerning the dispersion of Hg in the rivers of the MAMD date back to Dall'Aglio (1966) and Dall'Aglio et al.  
149 (1966), who detected extensive Hg contamination in stream sediments and waters. Bombace et al. (1973)  
150 estimated that at least 165 t of Hg were dispersed in the Paglia River from 1954 to 1963, the main period of  
151 mining activity. The same authors found up to 10.5 mg kg<sup>-1</sup> of Hg in stream sediments in the Paglia River, and  
152 up to 71.1 mg kg<sup>-1</sup> in the Siele creek, a right-side tributary (Fig. 1C). As stressed by Colica et al. (2019), the  
153 Paglia River overbank sediments represent a secondary pollution source, containing not less than 63 t of Hg.

154

155 *2.2 Geomorphological and multi-temporal analysis of channel changes*

156 The temporal changes of the Paglia riverbed and the extent of its alluvial deposits were reconstructed by GIS-  
157 based analysis of the available maps and aerial photos from the period of maximum mining production to date.  
158 Specifically, we used:

- 159 • Topographic maps produced by the IGM (Istituto Geografico Militare, Italy) dated 1883 (scale  
160 1:50,000), coinciding with the initial period of the MAMD mining activity.
- 161 • Aerial photos taken in 1954 (IGM, scale 1:33,000), coinciding with the maximum production period of  
162 MAMD (Caselli et al. 2007).
- 163 • Aerial photos taken in 1978 (IGM, scale 1:33,000), coinciding with the final production period of MAMD.
- 164 • Aerial photos from 1988 to 2016 (post-production period, during which partial reclamation of two main  
165 mining and smelting centers (Siele and Abbadia S.S.) was undertaken).
- 166 • Satellite images from Google Earth in 2019 (current status).

167 Through the open-source software Qgis 3.16 (Hannover; <https://qgis.org/it/site/>), the main morphological  
168 characters of the riverbed and floodplain were vectorized. The result consists of two vector layers: a linear  
169 type, representing the riverbed (dashed lines in the figures); and a polygonal one, corresponding to the  
170 floodplain (colored fill in the figures). Areas and widths were calculated by using the QGIS *Calculator Field*  
171 tool.

172 In Table 1 we report the definition of all the geomorphological terms used in the text.

173

174 *2.3. Soil and stream sediment sampling*

175 The geochemical analyses of soil (n = 74) and stream sediments (n = 17) presented in this study combine data  
176 of soil and stream sediment collected along five transects transverse to the Paglia River course (TP1, TP2,  
177 TP3, T4, TP5; Fig. 1C). New samples were collected from transects studied in previous works by Colica et al.  
178 (2019) (TP2 and TP4) and Rimondi et al. (2019) (TP1 and TP5), in the Tuscan portion of the Paglia River basin  
179 (Tab. 2).

180 A new transect (TP3) was chosen to integrate the previous ones. Stream sediments refer to the active Paglia  
181 River main course and were collected in the top layer (top 5-10 cm), below the water surface. Soils were  
182 collected about every 2 or 3 meters along the transect in the superficial horizon (0-30 cm) of the floodplain. All  
183 samples were collected as composite samples of about 1 kg, made up by mixing five sub-samples taken within  
184 a square of 5 m side around the selected sampling point, by using a shovel.

185

186 *2.4 Geochemical analysis*

187 The ARPAT laboratory (Siena, Italy) carried out sample preparation and chemical analysis of collected  
188 sediments and soils. Soils and sediments were homogenized, dried in air, sieved with a 2 mm sieve (as  
189 required by Italian national guidelines; D.Lgs. 152/2006), and then pulverized with a rotating ball mortar.  
190 Following the same Italian national guidelines, Hg concentrations were determined in the fraction <2mm, and  
191 were then recalculated to the whole samples (i.e., including the fraction >2 mm). This procedure is mandatory  
192 for Italian environmental agencies; in any case, the fraction >2mm was minimal in all collected samples,

193 therefore the application of this methodology had a negligible effect on the analytical results. Prior to analysis,  
194 soil and sediment powders were digested in aqua regia in a microwave oven (U.S. EPA 2007; 2014 methods).  
195 Concentrations of Hg were determined by ICP-MS (Inductively Coupled Plasma Mass Spectroscopy; UNI EN  
196 2016). The ARPAT laboratory is subjected to periodical quality checks by an independent organization  
197 (Accredia) according to the standard ISO/IEC 17025 and it takes part to the SNPA interlaboratory network for  
198 cross-checking. Specifically for Hg analyses, accuracy is determined employing the certified material ERM  
199 CC141 (certified Hg content:  $0.083 \pm 0.017$  mg kg<sup>-1</sup>; average of laboratory analyses:  $0.079 \pm 0.009$  mg kg<sup>-1</sup>).  
200 The overall analytical precision of the method is <10%, as determined by replicate analyses of different aliquots  
201 of the same bulk sample.

202

### 203 **3. Results**

#### 204 *3.1 Geomorphological changes along the first section of the Paglia floodplain*

205 Aerial photos and maps from 1883 to 2019 allowed to reconstruct the temporal changes of the riverbed and  
206 the floodplain along all the transects, as shown in Figure 2, whereas changes in land-use and geomorphologic  
207 features around each transect are reported in the supplementary material (Fig. S1-S5). In the following, we  
208 will analyze the temporal changes of the Paglia River floodplain, with reference to area variations. Changes in  
209 the floodplain width measured along transects were also considered; however, local features and/or  
210 fluctuations (e.g., due to climate variability) may affect the general processes controlling this parameter.

211 In the investigated segment of the Paglia River, the main changes observed during the 1883–2019 time frame  
212 include anthropogenic intervention and modifications in the principal road network, building of an industrial  
213 area that occupies part of the river valley, modifications of crop field extension, and other changes in land use  
214 (see supplementary materials for further details).

215 Figure 3 shows that a reduction of the floodplain area occurred from the end of 1800. The decrease was more  
216 pronounced between 1954 and 1978, with a reduction of almost two thirds (about 62%) of the total area (from  
217 2.8 to 0.9 km<sup>2</sup>). After 1978, which broadly corresponds to the end of mining activity, the floodplain area was  
218 subjected to fluctuations, with a relative increase in the period 1988-1998 followed by a progressive slow  
219 decrease lasting about 15 years and concluded in 2010. After this year until today, the area increased.  
220 Specifically, in the three years from 2010 to 2013, the floodplain area doubled its extension (Fig. 3).

221 On the other hand, in the period 1883-2010 the riverbed experienced a distinct narrowing of its width at all the  
222 five transects (Fig. 3; Fig. S1-S5), with a marked reduction occurred between 1954 and 1978. In the following  
223 period, the riverbed width remained more or less constant in the upper part of the river, from transect TP1 to  
224 TP3. On the contrary, in correspondence with transect TP5, we notice a progressive enlargement (47%) since  
225 2000, while after the confluence with the Senna Creek, at transect TP4, the width of the Paglia River increases,  
226 especially during or after major flood events (e.g., after the 2012 flood).

227

#### 228 *3.2 Mercury concentrations in stream sediments and floodplain soils*

229 Stream sediments and soils sampled along the transects in the Paglia River show highly variable Hg contents  
230 (from <0.2 mg/kg to 100 mg kg<sup>-1</sup>). In Figure 4 the spatial distribution of Hg in sediments and soils along each

231 transect is represented in association with their elevation and lithology. The full dataset is reported in Table  
232 S1.

233 The highest concentration of Hg in stream sediments ( $64 \text{ mg kg}^{-1}$ ) was recorded at TP1, while in soils ( $100 \text{ mg}$   
234  $\text{kg}^{-1}$ ) at transect TP5. Elevated Hg concentrations ( $1.7\text{-}6.7 \text{ mg kg}^{-1}$ ) were also found in fine sediments collected  
235 along transect TP5. These sediments were deposited by a flood event in December 2019, which occurred  
236 shortly before the sampling campaign (January-February 2020). This event led to the partial flooding of the  
237 field on the left side of the Paglia River.

238 In correspondence of the transects the Paglia River floodplain is almost entirely anomalous in Hg, i.e. with  
239 concentrations above the legal limit ( $1 \text{ mg kg}^{-1}$ ) defined by the Italian law for soil for residential and green area  
240 use (D.Lgs. 152/06), as shown in Figures 2 and 4. The anomaly boundary can be identified with the pre-  
241 anthropic fluvial terraces dated to the Pleistocene (Colica et al. 2019). These Pleistocene terraces are located  
242 at higher topographic levels with respect to more recent terraces formed during periods of anthropic activity  
243 (Colica et al. 2019). Nevertheless, Hg anomalies ( $> 1 \text{ mg kg}^{-1}$ ) are exceptionally found at high topographic  
244 altitude and, in some instances, over the Pleistocene terraces (e.g., in the transects TP4 and TP5), typically  
245 nearby roads and houses (Fig. 4).

246

#### 247 **4. Discussion**

##### 248 *4.1 Geomorphological river changes: anthropogenic and natural control and impact on Hg distribution in the* 249 *floodplain*

250 Braided rivers reflect the ongoing adjustment to fluctuating flow and sediment yield, under high sediment  
251 delivery conditions coupled with lower sediment throughput, due to a gentle slope (Piegay et al. 2006). Before  
252 the mid-1950s, the Paglia River was characterized by several anastomosing channels, river bars and islands,  
253 extending over a large area, as observed in the 1883 maps (Fig. S6; cf. Cencetti et al. 2017). This braided  
254 stage of the Paglia River coincided with the period during which Hg production, and thus waste production, at  
255 MAMD reached its maximum. Conceivably, sediments delivered by MAMD mining and metallurgical activities  
256 played an important role in shaping the changes of the Paglia River floodplain. Local miners report that,  
257 throughout the mine activity, mining and metallurgical wastes were discharged along the local waterways  
258 during rainy periods, and eventually were collected by the Paglia River. Consequently, peaks of Hg production  
259 significantly impacted sediment yields in the Paglia River. One of the main production peaks occurred during  
260 the first decades of 1900, driven by the increasing demand of Hg fulminate employed during the World War I.  
261 As reported by Caselli et al. (2007), during this period the MAMD overcame Almadén in Hg flask trading. After  
262 the economic crisis in 1930, production decreased, and maintained low during the World War II, since the  
263 district was heavily bombed. After the war, the Hg market, and thus MAMD, had a new important pulse due to  
264 the Korean war (Caselli et al. 2007), up to the mid-1960s; in the 1970s the Hg demand began a constant  
265 decrease down to a complete halt, with the consequent closure of the mines and plant production site in 1982.  
266 The actual mass of the mine wastes produced can be roughly estimated from the total amount of Hg produced  
267 ( $102,000$  tons), by the average Hg content of the *tout-venant* (generally less than 1 wt. %; Strappa 1977), and  
268 by the metallurgical recovery rate (about 80%; Benvenuti and Costagliola 2016). Based on this scenario, about  
269  $12 \cdot 10^6$  tons of mining/metallurgical wastes may have been produced in the MAMD, corresponding to  $6 \cdot 10^6 \text{ m}^3$

270 of sediments (average density:  $2 \text{ t m}^{-3}$ ), the same order of magnitude of the sediment volume presently stored  
271 in the fluvial terraces of this waterway (cf. Colica et al. 2019). These estimates suggest that during its braided  
272 stage, in the northern stretch of the Paglia River the sediment input was probably high, and significantly  
273 contributing to consolidate the braided stage of the river for the first half of 1900s.

274 In the 1954-1978 timespan, the Paglia floodplain area dramatically shrunk, dropping from  $2.6 \text{ km}^2$  to less than  
275  $1 \text{ km}^2$  (Fig. 3). The following change to a single channel led to a significant reduction of the floodplain area and  
276 produced a local incision of its original valley, leaving most Hg contaminated sediments in its terraces, located  
277 at a higher level with respect to the present-day watercourse.

278 The decrease of the Paglia River floodplain extension was one of the most intense ever recorded compared  
279 to floodplain reduction occurred in other Italian rivers (see “phase II” described by Surian et al. 2009). The  
280 Paglia River underwent an average reduction of the channel width of about 64% from 1883 to 1954, followed  
281 by a further reduction of about 70% from 1954 to 2012. As a result of floodplain narrowing, the Paglia River  
282 changed from a pre-1950 braided morphology to the present day wandering single-channel river with low  
283 sinuosity (Fig. 2; Fig. S1-S6).

284 Our study is consistent with the scenario depicted by Cencetti et al. (2017) in the southern stretch of the Paglia  
285 River from the Tuscan/Latium border to Orvieto (Fig. 1B), where incision of the Paglia riverbed was enhanced  
286 by the erosion of the old floodplain consequent to the increase of gravel mining into the riverbed and recovery  
287 of land for agriculture, which reduced supplies of sediment and caused a deficit in sediment transport (Cencetti  
288 et al. 2017; Colica et al. 2019). The tendency to riverbed incision is actually a common phenomenon observed  
289 in the same period in many other Italian and Mediterranean rivers (e.g., Brenta, Piave, Cellina, Tagliamento,  
290 and Torre Rivers in Italy; and Rambla de la Viuda in Spain), mainly steered by gravel mining (Surian and  
291 Rinaldi 2003; Aringoli et al. 2015; Cencetti et al. 2017; Calle et al. 2017; Dang et al. 2014). Gravel mining was  
292 intense in Italy starting from 1950s up to 1980s (Surian et al. 2009), driven by the post World War II economic  
293 expansion, and impacted river hydromorphologies, leading to scarcity of sediments, unbalanced river systems  
294 and modification of the long-term river morphodynamics, long after cessation of gravel mining of the riverbed  
295 (e.g., Calle et al. 2017).

296 The change from a braided to a single channel river had a profound consequence on the distribution of Hg  
297 contamination in the Paglia River basin. Due to the gradual deepening and narrowing of this single fluvial  
298 channel, Hg-contaminated sediments were deposited at higher topographically levels than the channel itself.  
299 One of the main consequences of this process led to a change in the transport/deposition cycle and to a  
300 tendential loss of mobility of the material deposited on the overbanks. Therefore, nowadays Paglia River  
301 contaminated sediments in the overbanks are no longer reached by the water flowing along the river channel,  
302 except during flood events. The extent of the overbank deposits impacted by Hg pollution, broadly corresponds  
303 to the floodplain built up by the river during the past century. More precisely, along the examined transects,  
304 the extension of Hg contaminated sediments roughly coincides with the 1954 floodplain.

305 Additionally, anthropogenic intervention may have contributed and still contributes to the unusual Hg  
306 contamination ( $\text{Hg} \geq 1 \text{ mg kg}^{-1}$ ; D.Lgs. 152/2006) in areas not subjected to the direct influence of the Paglia  
307 River and its tributaries, i.e., at higher elevations than those reached by the Paglia River during floods, and at  
308 a higher elevation than the terraces formed in the last century. This is observed almost systematically where

309 transects intercept streets or houses, such as near transects TP1 and TP5. Construction works such as road  
310 embankments or foundations of houses may indeed contain anomalously high values of Hg, because between  
311 1954 and 1978 it was common practice the use of mining and metallurgical waste as building material.  
312 Another contribution to the dispersion of Hg can be ascribed to the indirect effect of agricultural practices (soil  
313 amendments, irrigation, or artificial drainage), that may have caused the rearrangement and redistribution of  
314 superficial soil layers and associated Hg in fields located in the alluvial floodplain (e.g., Montagne et al. 2009).  
315 In summary, our study indicates that the effectiveness with which Hg-contaminated sediments were  
316 entrapped/stored along the Paglia River is probably the result of an incidental interplay between i) Hg mining,  
317 that fed the Paglia River floodplain with large amounts of Hg-contaminated sediments during its braided stage,  
318 and ii) the economic expansion of Italy after the World War II and the subsequent changes of the morphological  
319 features of Paglia River (due to gravel mining and other anthropogenic modifications), that enhanced the  
320 change to a single channel morphology of the Paglia River.

321

#### 322 *4.2 Implications of flood events on Hg distribution and resilience of the river system*

323 Local river morphology, sediment input and runoff, land uses and climate variability control fluvial dynamics  
324 (Schumm and Harvey 1999; Grabowski and Gurnell 2016, Marchamalo et al. 2016; Calle et al. 2017; Owens  
325 2020). After the closure of Hg mining, the spatial pattern of Hg downstream the Paglia River became a function  
326 of floods and high-water events rather than of Hg released to the river from mining activity (dashed black areas  
327 in Fig. 2). In the last 10 years, flood events occurred along the Paglia River caused the erosion of part of the  
328 previously built river terraces (Pattelli et al, 2014; Cencetti et al. 2017; Colica et al. 2019). During the 2012  
329 flood, in the lower section of the Paglia River (after the Siele creek confluence), the riverbed temporarily  
330 occupied part of the 1954 floodplain, reactivating several bars (as for example at TP5, Fig. S5). The  
331 incremented high erosion capacity caused an enlargement of the local river channel. Additional examples of  
332 the substantial changes on the width of the riverbed were observed after a flood event in December 2019,  
333 when the collapse of the riverbanks and part of the Cassia Road, about 3 km upstream of the TP2 transect,  
334 occurred.

335 The impact of floods on river morphology in the northern segment, highlighted in Figure 3, led to the increase  
336 of the Paglia floodplain area after the 2010 flood. A similar phenomenon was observed by Cencetti et al. (2017)  
337 in the southern stretch of the river, emphasizing that floods may partially restore the Paglia riverbed extension.  
338 These authors observed that by reactivating sediment supply, floods may restabilize channel morphology to  
339 near-reference conditions (i.e., pre-1954, pre-single channel), adjusting fluvial landforms as a response to the  
340 new hydrodynamic conditions (Simon 1989; Simon and Rinaldi 2006; Calle et al. 2017). Floods indeed play a  
341 crucial role in reshaping the patterns of pollutants dispersal, eroding, and transporting contaminants  
342 temporarily stored in channel and on overbanks to the floodplain (Coynel et al. 2007; Novakova et al. 2015;  
343 Ciszewski and Grygar 2016; Ponting et al. 2020). In river systems draining mining areas, storm and flood  
344 events have a significant control on the episodic transport of contaminants, and the impacts have been  
345 described in other Hg mining districts (e.g., Širca et al. 1999; Whyte et al. 2000; Springborn et al. 2011; Singer  
346 et al. 2013; McKee et al. 2017). During floods, enormous quantities of Hg-contaminated particulate are



347 mobilized because of higher runoff and the increased capacity of the stream to erode riverbanks. Following  
348 erosion, Hg transported as particulate suspended matter may increase up to 80-fold (Whyte et al. 2000).  
349 In the Paglia River, a distinct increase in Hg content was recorded immediately after the 2012 flood in stream  
350 sediments collected around transect TP1, with up to 905 mg kg<sup>-1</sup> of Hg, with respect to pre-flood values of 14  
351 mg kg<sup>-1</sup> (Pattelli et al. 2014). Similarly, after a flood event in 2019, mud deposited in the fields close to transects  
352 TP5 and TP1 was characterized by Hg content up to 6.7 mg kg<sup>-1</sup> and 34 mg kg<sup>-1</sup> respectively (Fig. 2A and E,  
353 dotted areas). This recurrent phenomenon is highlighted in Fig. 5, showing that high Hg pulses in stream  
354 sediments are recorded during or shortly after floods along the northern stretch of the Paglia River.  
355 Figure 5 shows, in addition, that a marked increase in Hg in stream sediments is observed in connection with  
356 the main flood events occurred since 2010. On the other hand, in the last years a decrease of Hg concentration  
357 has not occurred with increasing distance from the mine site of Abbadia San Salvatore, as could be expected  
358 by a “natural clean up” of the system.  
359 Under normal water flow conditions, Hg associated to the Paglia stream sediments is progressively washed  
360 away or diluted by a solid load that is not anomalous in Hg. The shifts between normal flow and flood events  
361 enhance the erosion of Hg-rich old (syn-mining) terraces, representing the actual overbanks in some part of  
362 the river, causing an alternance of low and high Hg contents along the riverbed. As described in Figure 5 and  
363 pointed out by Pattelli et al. (2014) for the 2012 flood, Hg pulses and floods are almost systematically in phase.  
364 The variability of metal dispersal associated to the effects of flood-sediments sorting and the mixing of  
365 particulate-associated pollutants, may result in changes of one to two orders of magnitude in metal content  
366 over distances of centimetres (Ciszewski and Grygar 2016). Therefore, overbank deposits and channel bars  
367 in the Paglia River represent a secondary source of Hg pollution, leading to the periodical transport of  
368 temporarily stored Hg-rich sediments to the river channel and to the floodplain. This phenomenon prevents a  
369 decrease of Hg concentration over time at least in short time (i.e., decades). Overbank sediments may indeed  
370 represent long-term storage for fine sediments with a residence time of the order of 10<sup>2</sup>–10<sup>3</sup> years (Grygar et  
371 al. 2016). The constant re-mobilization of contaminated material makes the Paglia River system not very  
372 resilient. A similar process is occurring in the Siele Creek, one of the largest Paglia River tributaries (Fornasaro  
373 et al. 2022).  
374 Since the contaminated area along the Paglia River almost corresponds with the area identified by the  
375 hydraulic hazard map of the Tiber River management basin plan (Trigile et al. 2018; Fig. 6), in the next future  
376 it is expected that further Hg mobilization will take place during flood events. The recent broadening of the  
377 Paglia River, started in 2010, coincided with an increase in monthly water discharge observed from 2003,  
378 consequent of a higher frequency of extreme flood events (Pattelli et al. 2014; Rimondi et al. 2014; Cencetti et  
379 al. 2017). These events will be predictably influenced by the variations of the precipitation regime because of  
380 climate change (van Vliet et al. 2013; Papalexidou and Montanari 2019). More precisely in southern Europe and  
381 in the Mediterranean region it is expected an overall drastic reduction in precipitation, more pronounced in  
382 summer (-25-30%; Castellari et al. 2014) Regional-scale model projections for Italy show indeed a significant  
383 temperature increases for the period 2070-2100 and a reduction in the number of days with little rain, and, by  
384 contrast, an increase of days with heavy rainfall (Castellari et al. 2014). Frequent drought periods characterized  
385 by long periods of low water flow, with modest or almost no solid transport, will alternate with intense rainy

386 periods or flash floods, concentrating solid transport in few short events. Consequently, climate variability could  
387 contribute to control the Hg distribution and overall mobility from MAMD and the Paglia River floodplain up to  
388 the Mediterranean Sea by the way of the Tiber River.

389 Our study provides useful information for management authorities to define precaution actions (such as  
390 limitations of sediment remobilization, river dredging, instream mining) and to identify conservation measures  
391 in this area (e.g., tree planting on overbanks, retention basins, thresholds and/or selective weirs). Further  
392 monitoring is necessary to ensure that the environmental quality of the river will not be altered by the spatial  
393 variability of Hg contaminated sediments distribution. The same strategies can be applied to similar rivers  
394 draining metal-contaminated areas that changed their morphology from braided to narrower channel, which in  
395 time are likely to act as continuous sources of contaminated particles deposited in their abandoned floodplains.  
396 On the other hand, the knowledge of distribution patterns of contaminated sediments is useful to address  
397 geomorphologic issues, as they can represent a tracer within the sediment system, providing a useful marker  
398 to the extent of sedimentation in a certain period. Furthermore, by tracking the movements, re-working, and  
399 removal of these contaminated sediments the role of floodplains as sediment storages can be established at  
400 different timescales.

401

## 402 **5. Conclusions**

403 The geomorphological and morphodynamic changes of the Paglia River, combined with anthropogenic  
404 activities occurred in the last century, controlled the spatial variability of Hg concentration in channel sediments  
405 and floodplain deposits of the northern stretch of the Paglia River, downstream the Monte Amiata Mining  
406 District. The distribution of Hg observed in the Paglia River floodplain resulted from the interplay of Hg mining,  
407 that fed the floodplain with large amounts of Hg-contaminated sediments during the braided stage (end of  
408 1800-mid-1950s) and the subsequent morphological changes of the river, following World War II (including  
409 gravel mining and other anthropogenic modifications), that led to the single-channel morphology of the Paglia  
410 River. After mine closure, a reduction of Hg concentration over time in river sediments did not occur, as it could  
411 be expected. Because of the braided narrowing morphology, the Paglia River enhanced the erosion of old syn-  
412 mining terraces, rich in Hg, and redistributed Hg contaminated sediments. Consequently, the process of  
413 transport/deposition did not allow a natural “clean up” of the river system since the closure of the mining sites.  
414 The temporal and spatial variability of Hg distribution is therefore principally associated with the fluvial  
415 geomorphological changes more than to anthropogenic activities.

416 At present, the main factor controlling Hg distribution in the next future is identified in climate variability,  
417 triggering erosion/deposition and redistribution of previously stored Hg contaminated overbank sediments and  
418 in the floodplain. In the Paglia River upper section, the alternation of normal flow conditions and flood events  
419 affects the geomorphology of the river course contributing to make overbank erosion a permanent secondary  
420 source of Hg. The expected intensification of extreme weather events (high rain events, intense floods),  
421 consequent of climate change, makes this area a Hg source of remarkable environmental concern at the local  
422 (Paglia River), regional (Tiber River), and Mediterranean scales in the future.

423

## 424 **Acknowledgments**

425 We thank Mario Paolieri (Università di Firenze) and members of the ARPAT staff for support in the field and in  
426 the laboratory. We acknowledge the useful comments by the Editor and two anonymous reviewers.

427

#### 428 **Funding and Conflicts of interests**

429 The research was funded by a specific agreement between ARPAT (responsible: C.F.) and Università degli  
430 Studi di Firenze (responsible: P.C.). The authors declare that they have no known competing financial interests  
431 or personal relationships that could have appeared to influence the work reported in this paper.

432

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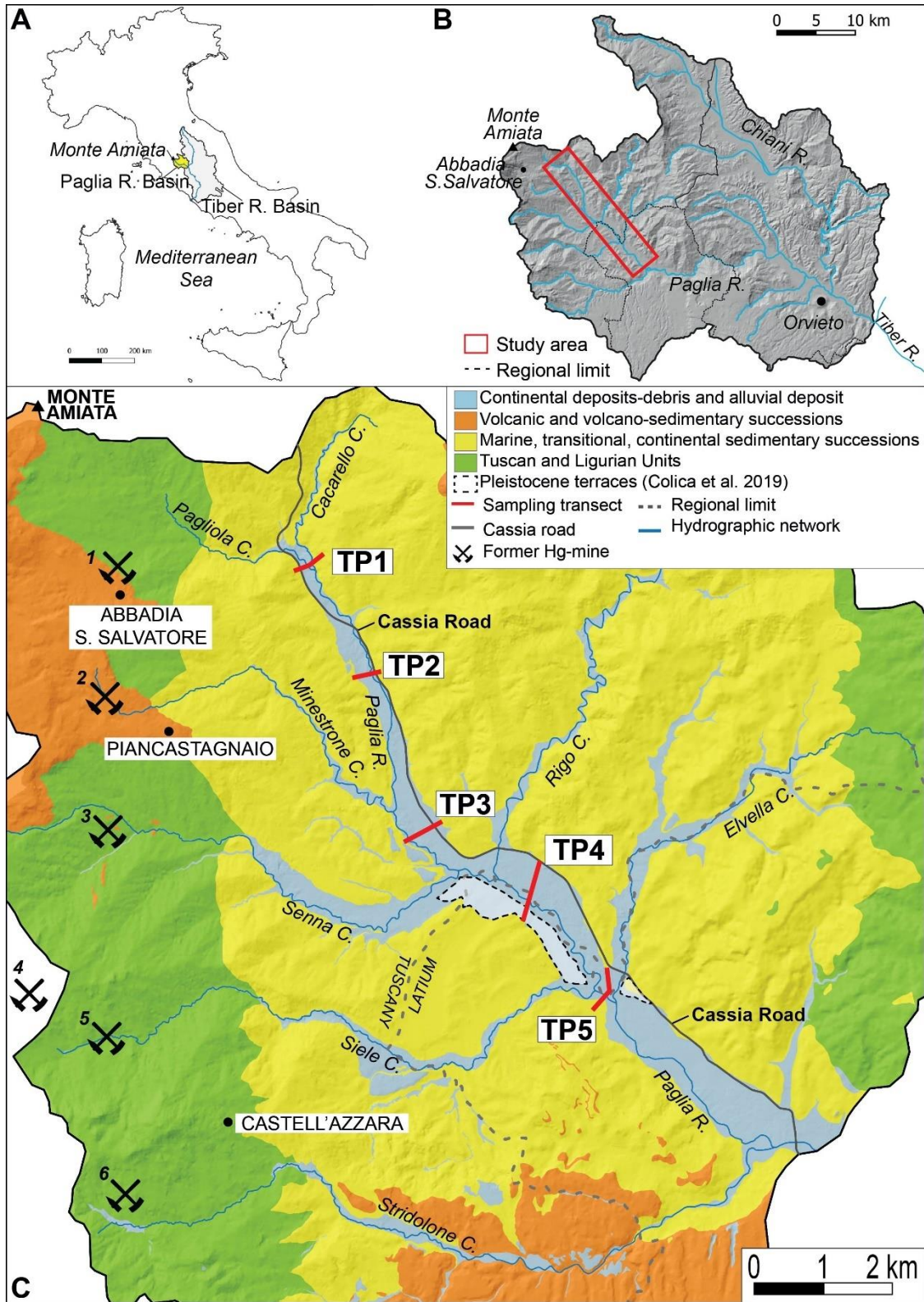
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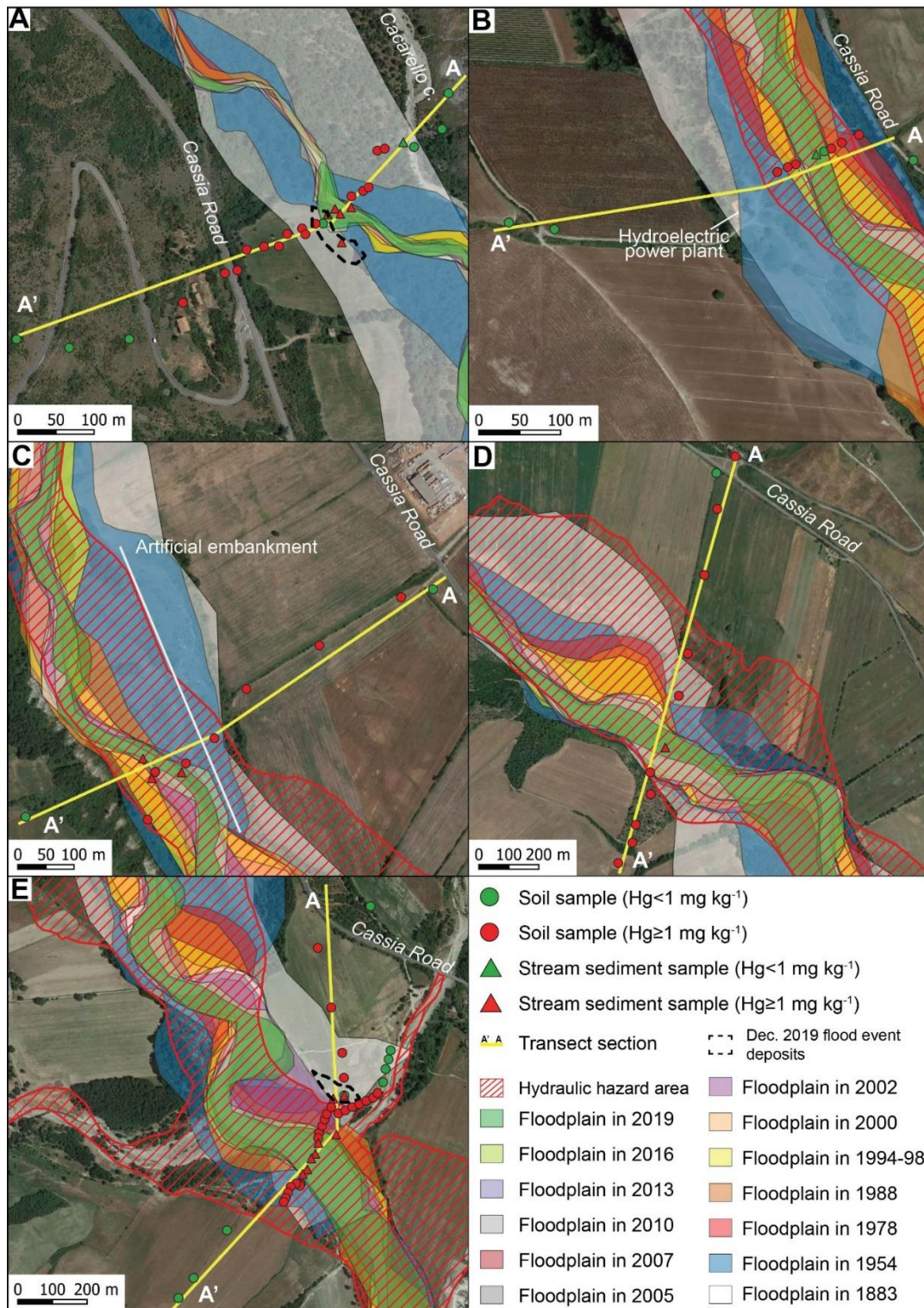
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<b>Element</b>	<b>Definition</b>	<b>Reference</b>
<b><i>Floodplain</i></b>	The floodplain is formed by past active channel riverbed abandonments. Two mechanisms, lateral migration by the braid-train and reactivation of abandoned channels within floodplains, operating separately or in combination, are responsible for floodplain reworking and their relatively young age (<250 years). Clearly, braided rivers can construct substantial areas of well-developed floodplain.	Aute et al. 2010
<b><i>River channel</i></b>	The active channel (or riverbed). The channel through which the water flows.	-
<b><i>Braided channel</i></b>	A network of channels formed in a river that has a great amount of sediment and a fluctuating pattern of discharge: the braiding effect is created by the formation of braid bars, around which the individual channels flow.	-
<b><i>Single channel with low sinuosity</i></b>	Sinuosity defines the degree of meandering of a riverbed. Channel sinuosity arises from flow hydraulic processes around bends in which secondary, across-channel circulation can increase meander wavelength and the migration of meanders across a floodplain. In general sinuosity is low in confined mountain streams.	Leopold et al. 1964
<b><i>Overbank sediment</i></b>	Overbank sediments occur along rivers and streams with variable water discharge. They are deposited on floodplains and levees from water suspension during floods, when the discharge exceeds the amounts that can be contained within the normal channel.	Bolviken et al. 2004
<b><i>Riverbank</i></b>	The landform distinguished by the topographic gradient from the bed of a channel along the lateral land-water margin up to the highest stage of flow or up to the topographic edge, where water begins to spread laterally over the floodplain surface.	Florsheim et al. 2008
<b><i>Bank erosion</i></b>	Bank erosion refers to the erosion of sediment from riverbank.	Florsheim et al. 2008

651 Tab. 1 Definition of the geomorphological terms used in the text.

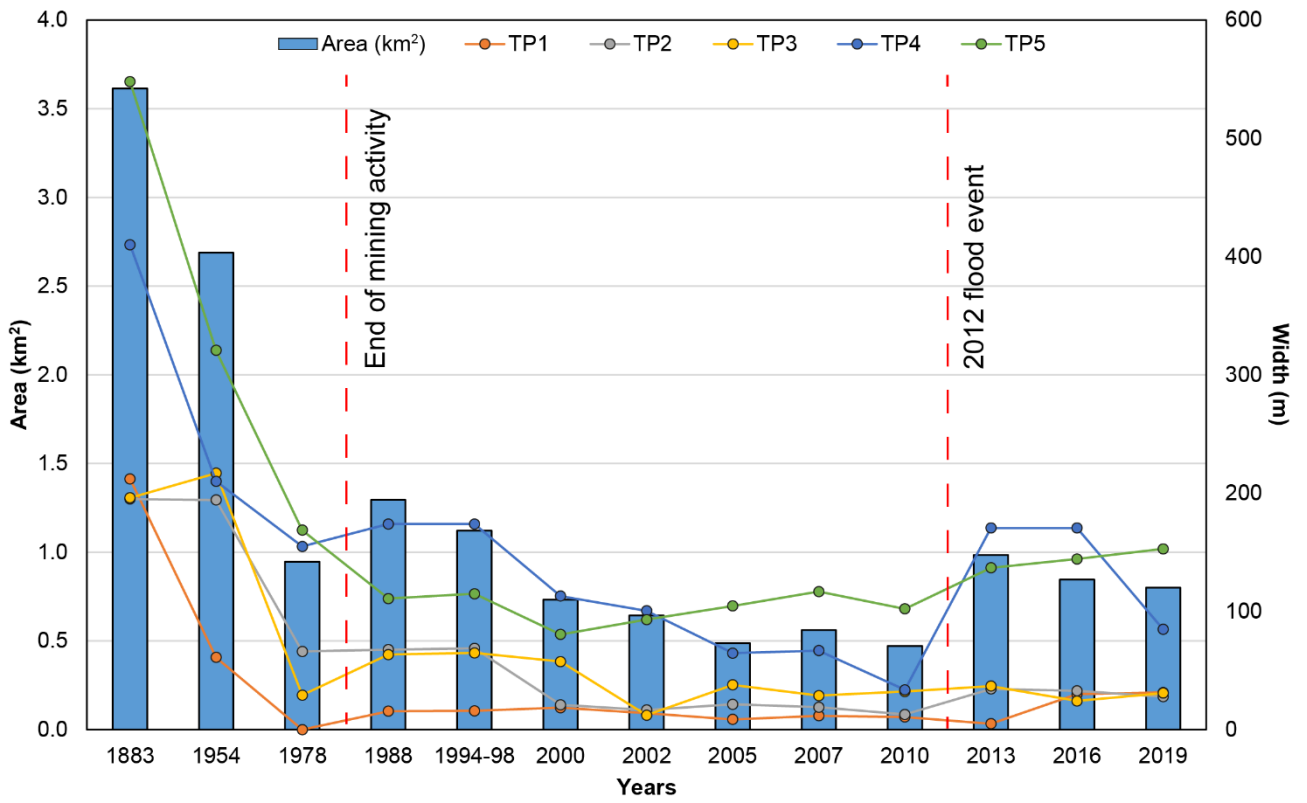


654  
 655 Fig. 1 A) Monte Amiata and Paglia-Tiber system location; B) Paglia River basin and its main tributaries; C)  
 656 Geological map of the upper part of the Paglia River basin. The location of the sampling transects and of the  
 657 main mines of MAMD are also reported: 1) Abbadia S. Salvatore; 2) Case di Paolo - Cerro della Tasca; 3)  
 658 Senna; 4) Solforate; 5) Siele; 6) Cornacchino.



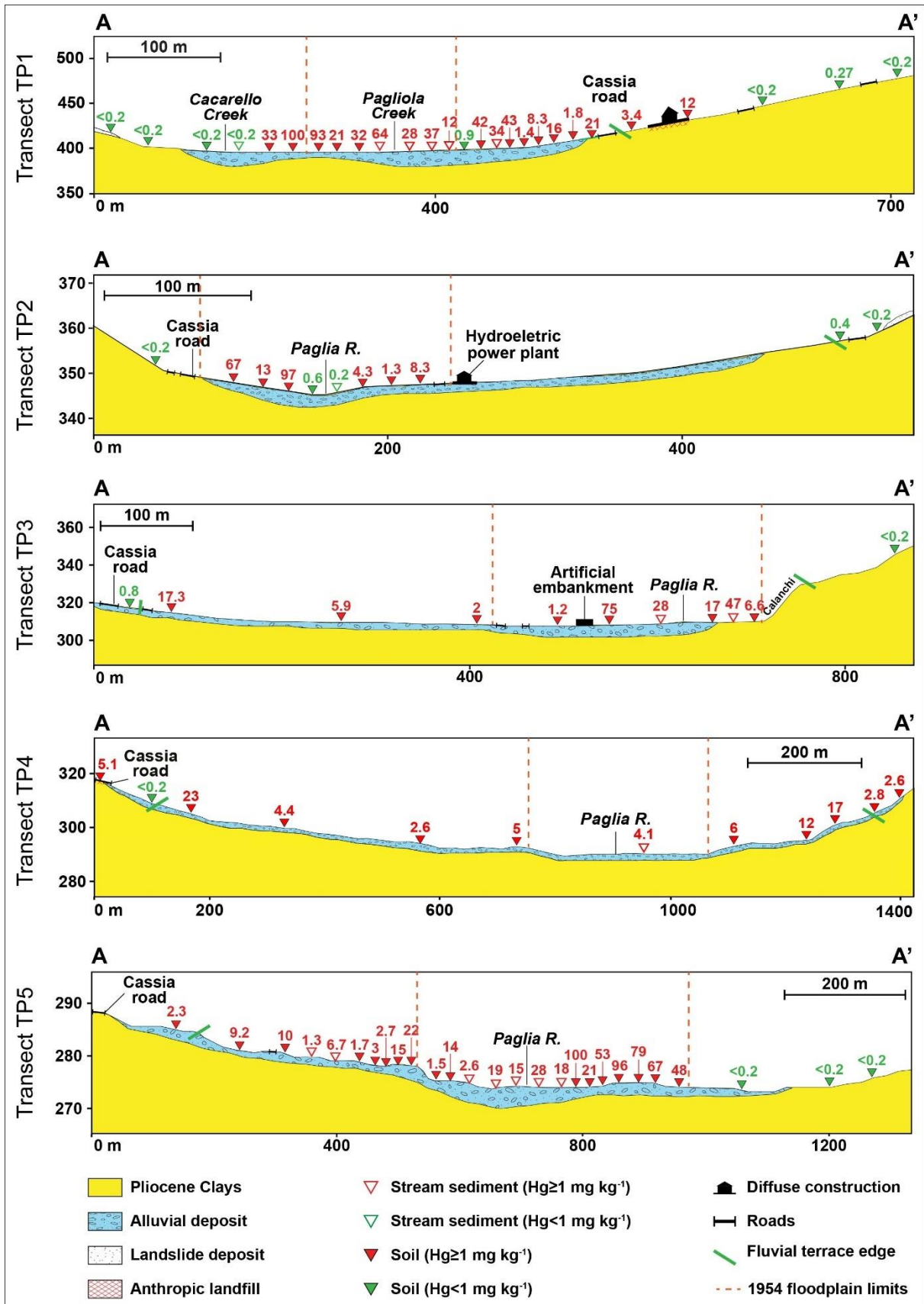
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660 Fig. 2 Satellite image (from Google Maps®, 2019) showing the samples collected along the transects in the  
 661 Paglia R. A) transect TP1; B) transect TP2; C) transect TP3; D) transect TP4; E) transect TP5. Samples with  
 662 Hg concentrations lower and higher than  $1 \text{ mg kg}^{-1}$  (D.Lgs. 152/2006) are shown in green and red, respectively.  
 663 The boundary of the riverbed (colored area) in different years from 1954 to 2019, the hydrologic hazard area  
 664 (lined area), the transects (yellow lines), and the December 2019 flood event deposits (blue dotted line in A  
 665 and E) are also shown.



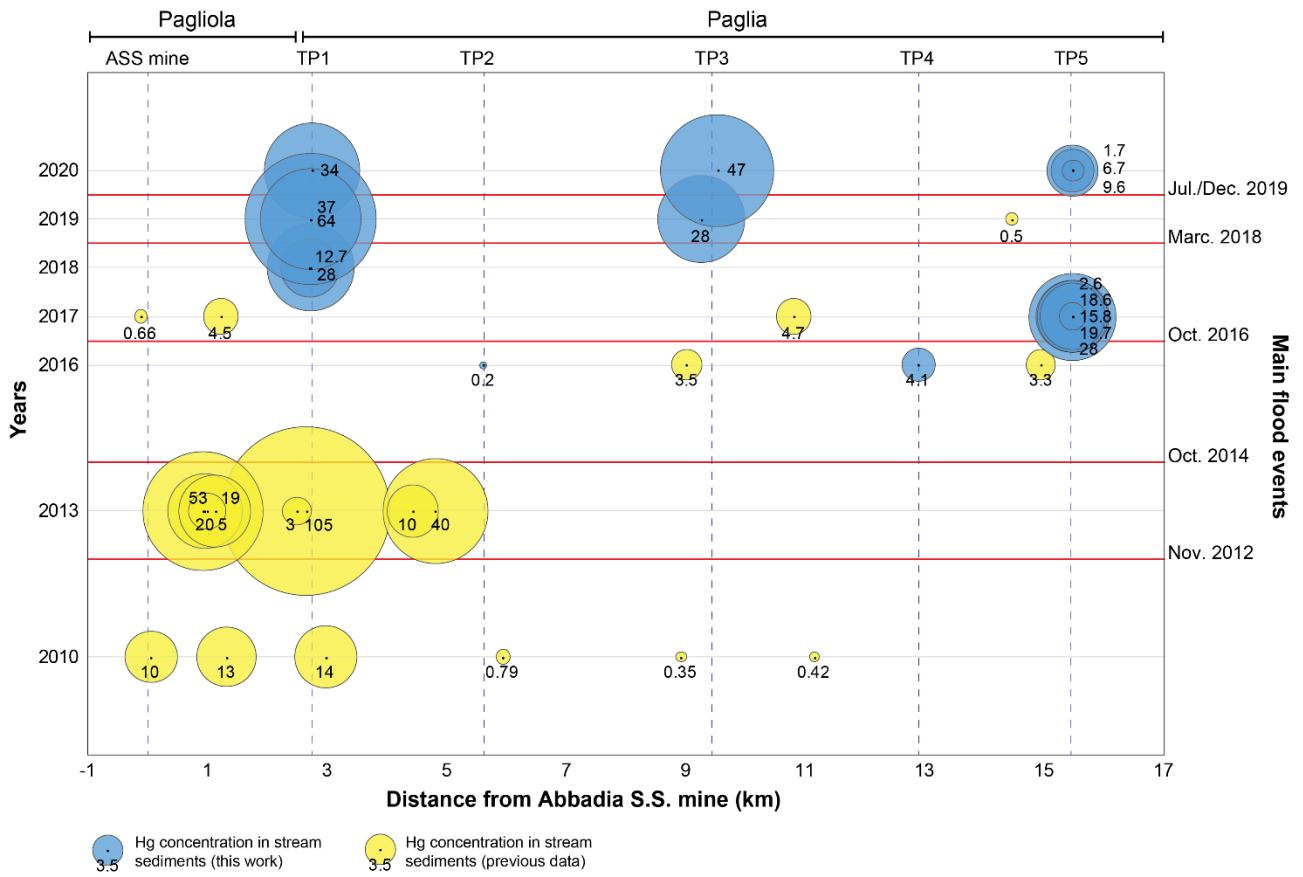
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667 Fig. 3 Floodplain area (m<sup>2</sup>) and width (m) along the sampled transects (colored lines) during the 1883-2019  
 668 period.



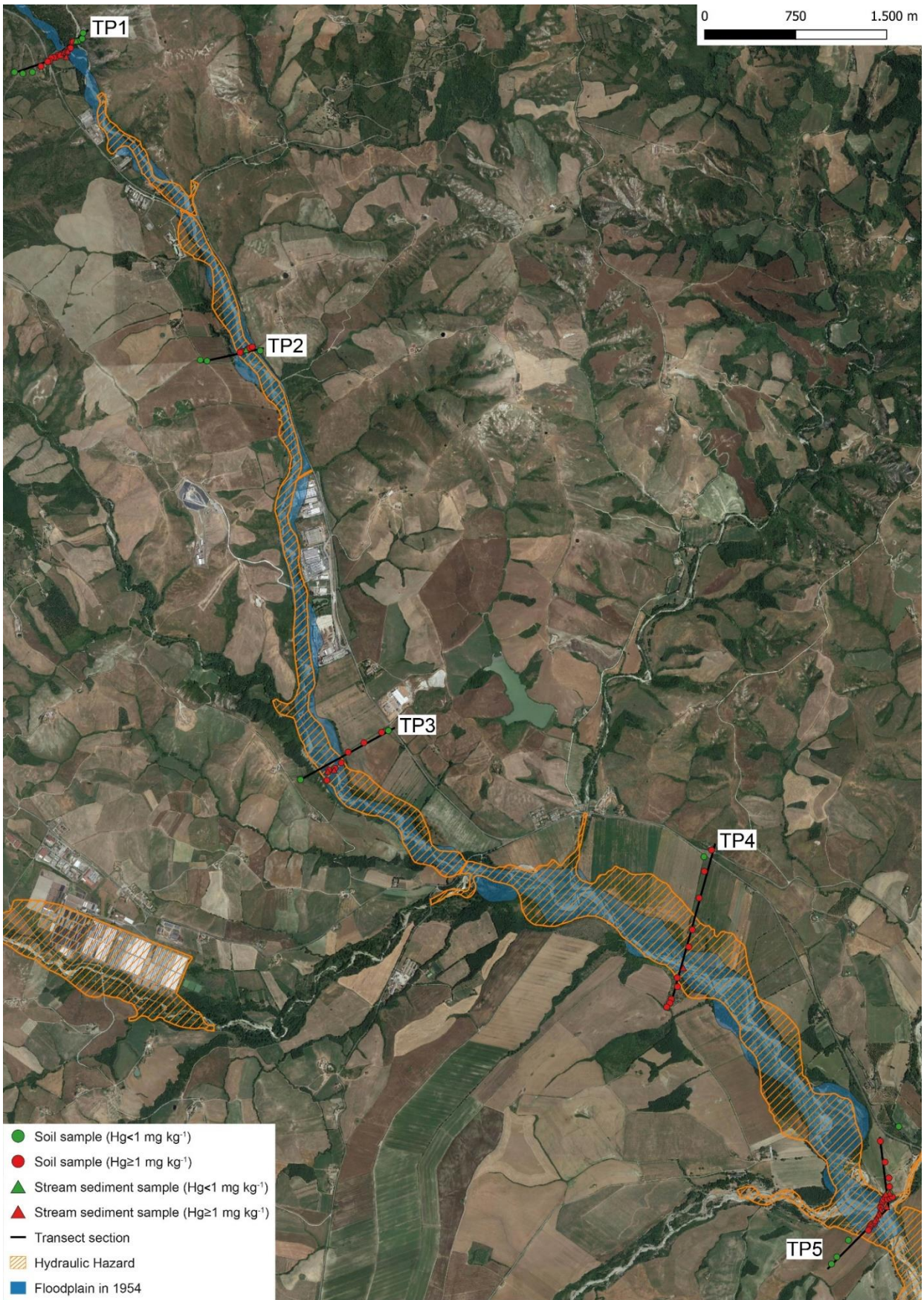
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670 Fig. 4 Geological sections, sample location, and Hg-concentration (mg/kg<sup>-1</sup>) in soils and sediments sampled  
 671 along the transects. View from North to South.



672

673 Fig. 5 Time-space variability of Hg concentrations (mg/kg<sup>-1</sup>) in stream sediments along the Paglia River course  
 674 in different years. Sampling location is indicated on the top X axis. The main flood events are also reported.





676 Fig. 6 Hydraulic hazard area (limits of Triglia et al., 2018) and Hg-concentration in stream and soil samples  
677 (mg kg<sup>-1</sup>).

1 **The extensive mercury contamination in soil and legacy sediments of the Paglia River basin (Tuscany,**  
2 **Italy): interplay between Hg-mining waste discharge along rivers, 1960s economic boom, and ongoing**  
3 **climate change**

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9

10 **Abstract**

11 **Purpose**

12 The extensive Hg contamination in soil and sediments occurring along the Paglia River (Central Italy) is the  
13 result of the interplay between the geomorphological changes of the river and the anthropic activities, primarily  
14 associated to the exploitation of Hg-deposits in the Monte Amiata mining district (MAMD). The present study  
15 points out the implications of the morphological changes occurred in the last 200 years of the Paglia River on  
16 the distribution of Hg along the floodplain and riverbed, which today represent one of the main Hg-reservoirs  
17 in the MAMD.

18 **Materials and Methods**

19 The temporal changes of the Paglia riverbed and the extent of its alluvial deposits were reconstructed by a  
20 GIS-based analysis of the available maps and aerial photos. The Hg-concentration in soil and sediment  
21 samples, collected along five transects transverse to the Paglia River course, was determined by ICP-MS.

22 **Results and Discussion**

23 Samples along the investigated Paglia River segment typically show Hg-contents exceeding the Italian  
24 threshold for residential and public green soil use (1 mg/kg<sup>-1</sup>).

25 The distribution of Hg in the Paglia floodplain results from the combination of exceedance of sediment yield to  
26 the river during mining activities, that fed the floodplain with large amounts of Hg-contaminated sediments  
27 during its braided stage about 100 years ago, and the morphological changes of the river, that led to the  
28 evolution from a braided to the present-day single channel river.

29 The magnitude of the extension of Hg-contamination, the river geomorphologic changes, and the processes  
30 of transport, deposition, and re-suspension did not allow a natural “clean up” of the river system, which shows  
31 a low resilience. Under high flow conditions, and especially in coincidence with intense rain events, large  
32 amounts of Hg stored in the overbanks sediments are mobilized and redistributed, contributing to make the  
33 floodplain a secondary Hg-source. Extreme weather events, expected to intensify as a consequence of climate  
34 change, will contribute to the recurrent distribution of Hg-contaminated legacy sediments in the floodplain and  
35 along the Paglia river course.

36 **Conclusion**

37 From a water/land management perspective, the variability of the river flow, associated with an increase of  
38 extreme flood events driven by climate change, will affect the distribution of Hg-contaminated particles in the  
39 Paglia River, contributing to the Hg input into the Mediterranean Sea in the future.

40

41 **Keywords**

42 legacy sediments; fluvial dynamics; mercury; Monte Amiata

43

44 **1. Introduction**

45 Geomorphic features of riverine systems result from the balance of many parameters (e.g., water and total  
46 sediment load; Schumm and Harvey 1999; Calle et al. 2017), that may in turn be affected by factors such as  
47 climate changes and human activities (Grabowski and Gurnell 2016; Marchamalo et al. 2016; Calle et al. 2017;  
48 Owens 2020; Vauclin et al. 2020). Specifically, in fluvial systems draining mining areas, mining activities may  
49 contribute significantly to the modification of river morphology, influencing sediment supply and the associated  
50 processes of erosion, transport, and deposition (e.g., Ciszewski and Grygar 2016; Davis et al. 2018). In  
51 addition, pollutants associated with mining particulate, such as heavy metals, are responsible of large-scale  
52 contamination up to several hundred kilometers away from the mining area (Martin and Maybeck 1979; Schafer  
53 et al. 2006; Mayes et al. 2013; Rimondi et al. 2019). Mining-contaminated legacy sediments deposited along  
54 waterways may remain stored within river channels and on floodplains for hundreds or thousands of years  
55 (Salomons and Förstner 1984; Macklin and Lewin 1989; Pavlowsky et al. 2017; Davis et al. 2018; Rimondi et  
56 al. 2019). They become diffuse sources of contamination if re-mobilized, for example by overbank erosion  
57 during flood events, or by human activities (e.g., gravel mining; Macklin et al. 1997; Pavlowsky et al. 2017;  
58 Colica et al. 2019). Floodplains therefore play an important role as both sinks and sources of metal  
59 contaminants in mined watersheds (Bradley 1989; Horowitz 1991; Lecce and Pavlowsky 1997; Coulthard and  
60 Macklin 2003; Ciszewski and Grygar 2016; Pavlowsky et al. 2017).

61 In the last 200 years, Italian waterways experienced considerable changes that triggered deep modifications  
62 of their original morphology (Surian and Rinaldi 2003; Cencetti et al. 2017, and references therein), similarly  
63 to other European rivers (e.g., Garcia-Ruiz et al. 2011; Debolini et al. 2015; Pavanelli et al. 2019). Incision and  
64 narrowing of the active channel were the most frequently observed modifications (Cencetti et al. 2017). The  
65 Tiber River (central Italy) and its tributary Paglia River (R.) are no exception and were affected by similar  
66 processes.

67 The Paglia River (49 km length) is one of the right-side tributaries of the Tiber River. Its morphological changes  
68 play a key role for the delivery of Hg to the Mediterranean Sea, since the river directly collects the runoff from  
69 one of the largest Hg ore districts in the world, the Monte Amiata Mining District (MAMD). Previous works  
70 described the morphological changes occurred in different sections of the Paglia River (e.g., Cencetti et al.  
71 2017), and the pervasive distribution of Hg in river sediments and soils of the Paglia River floodplain,  
72 highlighting how fluvial dynamics contribute to transport Hg contaminated sediments up to 200 km downstream  
73 the MAMD (Colica et al. 2019; Rimondi et al. 2019).

74 The present study points out the implications of the morphological changes in the last 200 years of the Paglia  
75 River (Tuscan stretch) on the buildup of the fluvial overbanks, which today represent one of the main Hg  
76 reservoirs in the MAMD district. Geochemical data obtained during a sampling campaign conducted in 2020  
77 by the Regional Environmental Protection Agency of the Tuscany (ARPAT) and by the environmental  
78 mineralogy group at Dipartimento di Scienze della Terra (DST), Università di Firenze, complement previous

79 studies by Colica et al. (2019) and Rimondi et al. (2019). These new results were integrated with previous data  
80 to assess the spatial and temporal variability of Hg contamination in the Paglia River floodplain.  
81 Specifically, the aims of this study are: i) to understand how the geomorphological (natural and anthropogenic)  
82 changes control the Hg distribution in the Paglia River floodplain; ii) to verify the implications of flood events  
83 on Hg distribution and resilience of the river system, **taking into account the potential consequences of climate**  
84 **changes.**

85

## 86 **2. Materials and methods**

### 87 *2.1 Study area*

#### 88 *2.1.1 Tiber-Paglia River system*

89 The Tiber River is the third longest river in Italy, flowing through the city of Rome and into the Mediterranean  
90 Sea (Cattuto et al. 1988; Ciccacci et al. 1988; Fredduzzi et al. 2007; Cencetti et al. 2017). The adjustments in  
91 its course in the last 250 ky were caused by an interplay between glacio-eustasy, sedimentary processes and  
92 regional uplifts (Marra et al. 2019). It was subjected to anthropic pressure probably before the establishment  
93 of the Roman Empire (Salomon et al. 2017). The most evident modifications took place during the 20th century  
94 and in the last decades following the Italian economic expansion, peaked between 1950 and 1970, with the  
95 construction of dams, sediment mining, and changes in the agriculture practices. The morphological changes  
96 occurred in the Tiber River are like those observed elsewhere in the Apennine area (Cencetti and Tacconi  
97 2005; Rinaldi and Simon 1998; Surian and Rinaldi 2003).

98 The Paglia River is one of the right-side tributaries of the Tiber River and arises from the junction of Pagliola  
99 an Cacarello creeks (388 m a.s.l., Fig. 1). The former drains the main mining and metallurgical center of MAMD  
100 near the Abbadia San Salvatore township. The Paglia River starts its course southeastward along a gentle  
101 slope, reaches the border between Tuscan and Latium regions (266 m a.s.l.) about 15 km from the starting  
102 point, and it enters into the Tiber River south of the city of Orvieto (Fig. 1C).

103 The geological-geomorphological structure of the upper basin of the Paglia River is linked to the formation of  
104 the Apennines during the Tertiary, and the subsequent post-collisional events (Marroni et al. 2015). The  
105 geology of the basin from the base to the top includes (Fig. 1C; Marroni et al. 2015): i) Tuscan and Ligurian  
106 Units (Paleozoic – Lower Miocene); ii) marine, transitional, and continental sedimentary successions (Lower  
107 Pliocene – Quaternary); iii) volcanic and volcano-sedimentary successions (Upper Pliocene – Upper  
108 Pleistocene); iv) continental deposits-debris and alluvial deposit (Quaternary). The quaternary continental  
109 deposits are characterized by i) holocenic fluvial deposits, present along the Paglia River valleys and its main  
110 tributaries, consisting mainly of sandy-silty beds and pebbles. The Paglia River cuts through its alluvial  
111 deposits, locally forming various orders of terraces; ii) Pleistocene deposits: fluvial-lacustrine deposits, mainly  
112 formed by conglomerates with sandy-silty beds levels. These deposits are arranged on large terraces located  
113 at higher elevation (from 5 to 20 m) compared to the current course of the Paglia River (e.g., Colica et al.  
114 2019). The presence of substrates characterized by erodible lithologies (Mio-Pliocene deposits) contributed to  
115 erosion processes with the formation of peculiar morphologies such as the "*biancane*" and the "*calanchi*"  
116 (Ciccacci et al. 2009). The shape of the Paglia River basin and the trend of the hydrographic network are

117 closely correlated with the structural characteristics of the Radicofani and Cetona grabens, set on normal fault  
118 systems with NNW-SSE trend, and trend-transforming systems with WSW-ENE trend (Sani et al. 2016).  
119 This area is characterized by a Mediterranean temperate climate, with hot and dry summer and cold and rainy  
120 winter. The average annual temperature is 10.5°C (period 1953 – 2000), and the average annual precipitation  
121 is 1480 mm (over the period 1925 – 2000; Ciccacci et al. 2009). About  $\frac{2}{3}$  of the total annual precipitation is  
122 concentrated in the autumn-winter season (Ciccacci et al. 1988). The Paglia River flow regime is controlled by  
123 seasonal variability, ranging from  $0.3 \text{ m}^3 \text{ s}^{-1}$  to  $26 \text{ m}^3 \text{ s}^{-1}$  monthly average (Fredduzzi 2005), with an annual  
124 average discharge of  $2.45 \text{ m}^3/\text{s}^{-1}$  (Cencetti et al. 2017). The hydrological periods are difficult to define (Moretti  
125 et al. 1988), due to the torrential regime of the initial part of the Paglia River, that may rapidly reach very high  
126 flow, collecting water contributes of a high number of tributaries. In general, the lowest and the highest water  
127 flow levels were recorded at the end of the summer period (September-October) and in winter-spring  
128 (November-March; Rimondi et al. 2014), respectively. An increase in mean monthly discharge was observed  
129 since 2003, due to a higher frequency of extreme flood events (Pattelli et al. 2014; Rimondi et al. 2014; Cencetti  
130 et al. 2017). This trend peaked with the flood of 2012 (mean monthly discharge:  $91.7 \text{ m}^3/\text{s}^{-1}$ , peak flow:  $2663$   
131  $\text{m}^3/\text{s}$ ; Cencetti et al. 2017).  
132 The Paglia River is physically shaped by sequential seasonal events of flooding and drying over a yearly cycle  
133 (Gasith and Resh 1999), reflecting the highly irregular rainfall patterns, with marked differences between wet  
134 and dry seasons, as most Mediterranean rivers. The concomitance of intense rainy days after dry summer  
135 periods, coupled with the scarce vegetation in the area, causes flash floods, with associated sliding-like mud  
136 and debris flows (Di Tria et al. 1999).

### 137 138 2.1.2 The Monte Amiata mining district (MAMD)

139 The MAMD district covers an area of  $\sim 400 \text{ km}^2$  and includes 42 former mines and 4 Hg roasting plants (Ferrara  
140 et al. 1998). The MAMD produced about 102,000 t of Hg between 1860s and 1980s (Colica et al. 2019),  
141 representing the third cumulative production ever reached in the world (Rimondi et al. 2015). The on-site  
142 metallurgical processing of cinnabar, the principal ore mineral, produced wastes, called calcines, with  
143 significant residual Hg contents ( $25 - 1500 \text{ mg/kg}^{-1}$ ; Rimondi et al. 2012; 2015). These wastes were often  
144 abandoned or discharged directly into the rivers adjacent to the mines, or used as filling material for road  
145 networks, house foundations, or landfills in crops (unpublished report, item T-1268 of the archives of the  
146 exploration company RIMIN). Numerous studies highlighted the environmental impact caused by over a  
147 century of mining and metallurgical activities in the MAMD, and the consequent contamination of Hg in the  
148 sediments transported by the Paglia River (Rimondi et al. 2019, and references therein). The first studies  
149 concerning the dispersion of Hg in the rivers of the MAMD date back to Dall'Aglio (1966) and Dall'Aglio et al.  
150 (1966), who detected extensive Hg contamination in stream sediments and waters. Bombace et al. (1973)  
151 estimated that at least 165 t of Hg were dispersed in the Paglia River from 1954 to 1963, the main period of  
152 mining activity. The same authors found up to  $10.5 \text{ mg/kg}^{-1}$  of Hg in stream sediments in the Paglia River, and  
153 up to  $71.1 \text{ mg/kg}^{-1}$  in the Siele creek, a right-side tributary (Fig. 1C). As stressed by Colica et al. (2019), the  
154 Paglia River overbanks sediments represent a secondary pollution source, containing not less than 63 t of Hg.  
155

156 *2.2 Geomorphological and multi-temporal analysis of channel changes*

157 The temporal changes of the Paglia riverbed and the extent of its alluvial deposits were reconstructed by GIS-  
158 based analysis of the available maps and aerial photos from the period of maximum mining production to date.  
159 Specifically, we used:

- 160 • Topographic maps produced by the IGM (Istituto Geografico Militare, Italy) dated 1883 (scale  
161 1:50,000), coinciding with the initial period of the MAMD mining activity.
- 162 • Aerial photos taken in 1954 (IGM, scale 1:33,000), coinciding with the maximum production period of  
163 MAMD (Caselli et al. 2007).
- 164 • Aerial photos taken in 1978 (IGM, scale 1:33,000), coinciding with the final production period of MAMD.
- 165 • Aerial photos from 1988 to 2016 (post-production period, during which partial reclamation of two main  
166 mining and smelting centers (Siele and Abbadia S.S.) was undertaken).
- 167 • Satellite images from Google Earth in 2019 (current status).

168 Through the open-source software Qgis 3.16 (Hannover; <https://qgis.org/it/site/>), the main morphological  
169 characters of the riverbed and floodplain were vectorized. The result consists of two vector layers: a linear  
170 type, representing the riverbed (dashed lines in the figures); and a polygonal one, corresponding to the  
171 floodplain (colored fill in the figures). Areas and widths were calculated by using the QGIS *Calculator Field*  
172 tool.

173 In Table 1 we report the definition of all the geomorphological terms used in the text.

174

175 *2.3. Soil and stream sediment sampling*

176 The geochemical analyses of soil (n = 74) and stream sediments (n = 17) presented in this study combine data  
177 of soil and stream sediment collected along five transects transverse to the Paglia River course (TP1, TP2,  
178 TP3, T4, TP5; Fig. 1C). New samples were collected from transects studied in previous works by Colica et al.  
179 (2019) (TP2 and TP4) and Rimondi et al. (2019) (TP1 and TP5), in the Tuscan portion of the Paglia River basin  
180 (Tab. 2).

181 A new transect (TP3) was chosen to integrate the previous ones. Stream sediments refer to the active Paglia  
182 River main course and were collected in the top layer (top 5-10 cm), below the water surface. Soils were  
183 collected about every 2 or 3 meters along the transect in the superficial horizon (0-30 cm) of the floodplain. All  
184 samples were collected as composite samples of about 1 kg, made up by mixing five sub-samples taken within  
185 a square of 5 m side around the selected sampling point, by using a shovel.

186

187 *2.4 Geochemical analysis*

188 The ARPAT laboratory (Siena, Italy) carried out sample preparation and chemical analysis of collected  
189 sediments and soils. Soils and sediments were homogenized, dried in air, sieved with a 2 mm sieve (as  
190 required by Italian national guidelines; D.Lgs. 152/2006), and then pulverized with a rotating ball mortar.  
191 Following the same Italian national guidelines, Hg concentrations were determined in the fraction <2mm, and  
192 were then recalculated to the whole samples (i.e., including the fraction >2 mm). This procedure is mandatory  
193 for Italian environmental agencies; in any case, the fraction >2mm was minimal in all collected samples,

194 therefore the application of this methodology had a negligible effect on the analytical results. Prior to analysis,  
195 soil and sediment powders were digested in aqua regia in a microwave oven (U.S. EPA 2007; 2014 methods).  
196 ~~“Water soluble Hg species” in sediment and soil samples were quantified with leaching tests. These were~~  
197 ~~conducted in deionized water using a soil/sediment to water ratio of 1:10, for a batch reaction time of 24 h,~~  
198 ~~following the methods UNI EN (UNI EN 2004). Since the results denote low water soluble Hg species, these~~  
199 ~~data were not presented in the following.~~

200 Concentrations of Hg were determined by ICP-MS (Inductively Coupled Plasma Mass Spectroscopy; UNI EN  
201 2016). The ARPAT laboratory is subjected to periodical quality checks by an independent organization  
202 (Accredia) according to the standard ISO/IEC 17025 and it takes part to the SNPA interlaboratory network for  
203 cross-checking. Specifically for Hg analyses, accuracy is determined employing the certified material ERM  
204 CC141 (certified Hg content:  $0.083 \pm 0.017$  mg kg<sup>-1</sup>; average of laboratory analyses:  $0.079 \pm 0.009$  mg kg<sup>-1</sup>).  
205 The overall analytical precision of the method is <10%, as determined by replicate analyses of different aliquots  
206 of the same bulk sample.

207

### 208 3. Results

#### 209 3.1 Geomorphological changes along the first section of the Paglia floodplain

210 Aerial photos and maps from 1883 to 2019 allowed to reconstruct the temporal changes of the riverbed and  
211 the floodplain along all the transects, as shown in Figure 2, whereas changes in land-use and geomorphologic  
212 features around each transect are reported in the supplementary material (Fig. S1-S5). In the following, we  
213 will analyze the temporal changes of the Paglia River floodplain, with reference to area variations. Changes in  
214 the floodplain width measured along transects were also considered; however, local features and/or  
215 fluctuations (e.g., due to climate variability) may affect the general processes controlling this parameter.

216 In the investigated segment of the Paglia River, the main changes observed during the 1883–2019 time frame  
217 include anthropogenic intervention and modifications in the principal road network, building of an industrial  
218 area that occupies part of the river valley, modifications of crop field extension, and other changes in land use  
219 (see supplementary materials for further details).

220 Figure 3 shows that a reduction of the floodplain area occurred from the end of 1800. The decrease was more  
221 pronounced between 1954 and 1978, with a reduction of almost two thirds (about 62%) of the total area (from  
222 2.8 to 0.9 km<sup>2</sup>). After 1978, which broadly corresponds to the end of mining activity, the floodplain area was  
223 subjected to fluctuations, with a relative increase in the period 1988-1998 followed by a progressive slow  
224 decrease lasting about 15 years and concluded in 2010. After this year until today, the area increased.  
225 Specifically, in the three years from 2010 to 2013, the floodplain area doubled its extension (Fig. 3).

226 On the other hand, in the period 1883-2010 the riverbed experienced a distinct narrowing of its width at all the  
227 five transects (Fig. 3; Fig. S1-S5), with a marked reduction occurred between 1954 and 1978. In the following  
228 period, the riverbed width remained more or less constant in the upper part of the river, from transect TP1 to  
229 TP3. On the contrary, in correspondence with transect TP5, we notice a progressive enlargement (47%) since  
230 2000, while after the confluence with the Senna Creek, at transect TP4, the width of the Paglia River increases,  
231 especially during or after major flood events (e.g., after the 2012 flood).

232

233 *3.2 Mercury concentrations in stream sediments and floodplain soils*

234 Stream sediments and soils sampled along the transects in the Paglia River show highly variable Hg contents  
235 (from <0.2 mg/kg to 100 mg/kg<sup>-1</sup>). In Figure 4 the spatial distribution of Hg in sediments and soils along each  
236 transect is represented in association with their elevation and lithology. The full dataset is reported in Table  
237 S1.

238 The highest concentration of Hg in stream sediments (64 mg/kg<sup>-1</sup>) was recorded at TP1, while in soils (100  
239 mg/kg<sup>-1</sup>) at transect TP5. Elevated Hg concentrations (1.7-6.7 mg/kg<sup>-1</sup>) were also found in fine sediments  
240 collected along transect TP5. These sediments were deposited by a flood event in December 2019, which  
241 occurred shortly before the sampling campaign (January-February 2020). This event led to the partial flooding  
242 of the field on the left side of the Paglia River.

243 In correspondence of the transects the Paglia River floodplain is almost entirely anomalous in Hg, i.e. with  
244 concentrations above the legal limit (1 mg/kg<sup>-1</sup>) defined by the Italian law for soil for residential and green area  
245 use (D.Lgs. 152/06), as shown in Figures 2 and 4. The anomaly boundary can be identified with the pre-  
246 anthropic fluvial terraces dated to the Pleistocene (Colica et al. 2019). These Pleistocene terraces are located  
247 at higher topographic levels with respect to more recent terraces formed during periods of anthropic activity  
248 (Colica et al. 2019). Nevertheless, Hg anomalies (> 1 mg/kg<sup>-1</sup>) are exceptionally found at high topographic  
249 altitude and, in some instances, over the Pleistocene terraces (e.g., in the transects TP4 and TP5), typically  
250 nearby roads and houses (Fig. 4).

251

252 **4. Discussion**

253 *4.1 Geomorphological river changes: anthropogenic and natural control and impact on Hg distribution in the*  
254 *floodplain*

255 Braided rivers reflect the ongoing adjustment to fluctuating flow and sediment yield, under high sediment  
256 delivery conditions coupled with lower sediment throughput, due to a gentle slope (Piegay et al. 2006). Before  
257 the mid-1950s, the Paglia River was characterized by several anastomosing channels, river bars and islands,  
258 extending over a large area, as observed in the 1883 maps (Fig. S6; cf. Cencetti et al. 2017). This braided  
259 stage of the Paglia River coincided with the period during which Hg production, and thus waste production, at  
260 MAMD reached its maximum. Conceivably, sediments delivered by MAMD mining and metallurgical activities  
261 played an important role in shaping the changes of the Paglia River floodplain. Local miners report that,  
262 throughout the mine activity, mining and metallurgical wastes were discharged along the local waterways  
263 during rainy periods, and eventually were collected by the Paglia River. Consequently, peaks of Hg production  
264 significantly impacted sediment yields in the Paglia River. One of the main production peaks occurred during  
265 the first decades of 1900, driven by the increasing demand of Hg fulminate employed during the I World War  
266 I. As reported by Caselli et al. (2007), during this period the MAMD overcame Almadén in Hg flask trading.  
267 After the economic crisis in 1930, production decreased, and maintained low during the II World War II, since  
268 the district was heavily bombed. After the war, the Hg market, and thus MAMD, had a new important pulse  
269 due to the Korean war (Caselli et al. 2007), up to the mid-1960s; in the 1970s the Hg demand began a constant  
270 decrease down to a complete halt, with the consequent closure of the mines and plant production site in 1982.



271 The actual mass of the mine wastes produced can be roughly estimated from the total amount of Hg produced  
272 (102,000 tons), by the average Hg content of the *tout-venant* (generally less than 1 wt. %; Strappa 1977), and  
273 by the metallurgical recovery rate (about 80%; Benvenuti and Costagliola 2016). Based on this scenario, about  
274  $12 \cdot 10^6$  tons of mining/metallurgical wastes may have been produced in the MAMD, corresponding to  $6 \cdot 10^6$  m<sup>3</sup>  
275 of sediments (average density: 2 t/m<sup>3</sup>), the same order of magnitude of the sediment volume presently stored  
276 in the fluvial terraces of this waterway (cf. Colica et al. 2019). These estimates suggest that during its braided  
277 stage, in the northern stretch of the Paglia River the sediment input was probably high, and significantly  
278 contributing to consolidate the braided stage of the river for the first half of 1900s.

279 In the 1954-1978 timespan, the Paglia floodplain area dramatically shrunk, dropping from 2.6 km<sup>2</sup> to less than  
280 1 km<sup>2</sup> (Fig. 3). The following change to a single channel led to a significant reduction of the floodplain area and  
281 produced a local incision of its original valley, leaving most Hg contaminated sediments in its terraces, located  
282 at a higher level with respect to the present-day watercourse.

283 The decrease of the Paglia River floodplain extension was one of the most intense ever recorded compared  
284 to floodplain reduction occurred in other Italian rivers (see “phase II” described by Surian et al. 2009). The  
285 Paglia River underwent an average reduction of the channel width of about 64% from 1883 to 1954, followed  
286 by a further reduction of about 70% from 1954 to 2012. As a result of floodplain narrowing, the Paglia River  
287 changed from a pre-1950 braided morphology to the present day wandering single-channel river with low  
288 sinuosity (Fig. 2; Fig. S1-S6).

289 Our study is consistent with the scenario depicted by Cencetti et al. (2017) in the southern stretch of the Paglia  
290 River from the Tuscan/Latium border to Orvieto (Fig. 1B), where incision of the Paglia riverbed was enhanced  
291 by the erosion of the old floodplain consequent to the increase of gravel mining into the riverbed and recovery  
292 of land for agriculture, which reduced supplies of sediment and caused a deficit in sediment transport (Cencetti  
293 et al. 2017; Colica et al. 2019). The tendency to riverbed incision is actually a common phenomenon observed  
294 in the same period in many other Italian and Mediterranean rivers (e.g., Brenta, Piave, Cellina, Tagliamento,  
295 and Torre Rivers in Italy; and Rambla de la Viuda in Spain), mainly steered by gravel mining (Surian and  
296 Rinaldi 2003; Aringoli et al. 2015; Cencetti et al. 2017; Calle et al. 2017; Dang et al. 2014). Gravel mining was  
297 intense in Italy starting from 1950s up to 1980s (Surian et al. 2009), driven by the post World War II economic  
298 expansion, and impacted river hydromorphologies, leading to scarcity of sediments, unbalanced river systems  
299 and modification of the long-term river morphodynamics, long after cessation of gravel mining of the riverbed  
300 (e.g., Calle et al. 2017).

301 The change from a braided to a single channel river had a profound consequence on the distribution of Hg  
302 contamination in the Paglia River basin. ~~Gradual deepening and narrowing of this single fluvial channel left~~  
303 ~~Hg-contaminated sedimentary bodies suspended at higher levels than the channel itself.~~ ~~Due to the gradual~~  
304 ~~deepening and narrowing of this single fluvial channel, Hg-contaminated sediments were deposited at higher~~  
305 ~~topographically levels than the channel itself.~~ One of the main consequences of this process led to a change  
306 in the transport/deposition cycle and to a tendential loss of mobility of the material deposited on the overbanks.  
307 Therefore, nowadays Paglia River contaminated sediments in the overbanks are no longer reached by the  
308 water flowing along the river channel, except during flood events. The extent of the overbank deposits impacted  
309 by Hg pollution, broadly corresponds to the floodplain built up by the river during the past century. More

310 precisely, along the examined transects, the extension of Hg contaminated sediments roughly coincides with  
311 the 1954 floodplain.

312 Additionally, anthropogenic intervention may have contributed and still contributes to the unusual Hg  
313 contamination ( $\text{Hg} \geq 1 \text{ mg/kg}^{-1}$ ; D.Lgs. 152/2006) in areas not subjected to the direct influence of the Paglia  
314 River and its tributaries, i.e., at higher elevations than those reached by the Paglia River during floods, and at  
315 a higher elevation than the terraces formed in the last century. This is observed almost systematically where  
316 transects intercept streets or houses, such as near transects TP1 and TP5. Construction works such as road  
317 embankments or foundations of houses may indeed contain anomalously high values of Hg, because between  
318 1954 and 1978 it was common practice the use of mining and metallurgical waste as building material.

319 Another contribution to the dispersion of Hg can be ascribed to the indirect effect of agricultural practices (soil  
320 amendments, irrigation, or artificial drainage), that may have caused the rearrangement and redistribution of  
321 superficial soil layers and associated Hg in fields located in the alluvial floodplain (e.g., Montagne et al. 2009).  
322 In summary, our study indicates that the effectiveness with which Hg-contaminated sediments were  
323 entrapped/stored along the Paglia River is probably the result of an incidental interplay between i) Hg mining,  
324 that fed the Paglia River floodplain with large amounts of Hg-contaminated sediments during its braided stage,  
325 and ii) the economic expansion of Italy after the **II World War II** and the subsequent changes of the  
326 morphological features of Paglia River (due to gravel mining and other anthropogenic modifications), that  
327 enhanced the change to a single channel morphology of the Paglia River.

328

#### 329 *4.2 Implications of flood events on Hg distribution and resilience of the river system*

330 Local river morphology, sediment input and runoff, land uses and climate variability control fluvial dynamics  
331 (Schumm and Harvey 1999; Grabowski and Gurnell 2016, Marchamalo et al. 2016; Calle et al. 2017; Owens  
332 2020). After the closure of Hg mining, the spatial pattern of Hg downstream the Paglia River became a function  
333 of floods and high-water events rather than of Hg released to the river from mining activity (dashed black areas  
334 in Fig. 2). In the last 10 years, flood events occurred along the Paglia River caused the erosion of part of the  
335 previously built river terraces (Pattelli et al, 2014; Cencetti et al. 2017; Colica et al. 2019). During the 2012  
336 flood, in the lower section of the Paglia River (after the Siele creek confluence), the riverbed temporarily  
337 occupied part of the 1954 floodplain, reactivating several bars (as for example at TP5, Fig. S5). The  
338 incremented high erosion capacity caused an enlargement of the local river channel. Additional examples of  
339 the substantial changes on the width of the riverbed were observed after a flood event in December 2019,  
340 when the collapse of the riverbanks and part of the Cassia Road, about 3 km upstream of the TP2 transect,  
341 occurred.

342 The impact of floods on river morphology in the northern segment, highlighted in Figure 3, led to the increase  
343 of the Paglia floodplain area after the 2010 flood. A similar phenomenon was observed by Cencetti et al. (2017)  
344 in the southern stretch of the river, emphasizing that floods may partially restore the Paglia riverbed extension.  
345 These authors observed that by reactivating sediment supply, floods may restabilize channel morphology to  
346 near-reference conditions (i.e., pre-1954, pre-single channel), adjusting fluvial landforms as a response to the  
347 new hydrodynamic conditions (Simon 1989; Simon and Rinaldi 2006; Calle et al. 2017). Floods indeed play a  
348 crucial role in reshaping the patterns of pollutants dispersal, eroding, and transporting contaminants

349 temporarily stored in channel and on overbanks to the floodplain (Coynel et al. 2007; Novakova et al. 2015;  
350 Ciszewski and Grygar 2016; Ponting et al. 2020). In river systems draining mining areas, storm and flood  
351 events have a significant control on the episodic transport of contaminants, and the impacts have been  
352 described in other Hg mining districts (e.g., Širca et al. 1999; Whyte et al. 2000; Springborn et al. 2011; Singer  
353 et al. 2013; McKee et al. 2017). During floods, enormous quantities of Hg-contaminated particulate are  
354 mobilized because of higher runoff and the increased capacity of the stream to erode riverbanks. Following  
355 erosion, Hg transported as particulate suspended matter may increase up to 80-fold (Whyte et al. 2000).  
356 In the Paglia River, a distinct increase in Hg content was recorded immediately after the 2012 flood in stream  
357 sediments collected around transect TP1, with up to 905 mg/kg<sup>-1</sup> of Hg, with respect to pre-flood values of 14  
358 mg/kg<sup>-1</sup> (Pattelli et al. 2014). Similarly, after a flood event in 2019, mud deposited in the fields close to transects  
359 TP5 and TP1 was characterized by Hg content up to 6.7 mg/kg<sup>-1</sup> and 34 mg/kg<sup>-1</sup> respectively (Fig. 2A and E,  
360 dotted areas). This recurrent phenomenon is highlighted in Fig. 5, showing that high Hg pulses in stream  
361 sediments are recorded during or shortly after floods along the northern stretch of the Paglia River.  
362 Figure 5 shows, in addition, that a marked increase in Hg in stream sediments is observed in connection with  
363 the main flood events occurred since 2010. On the other hand, in the last years a decrease of Hg concentration  
364 has not occurred with increasing distance from the mine site of Abbadia San Salvatore, as could be expected  
365 by a “natural clean up” of the system.  
366 Under normal water flow conditions, Hg associated to the Paglia stream sediments is progressively washed  
367 away or diluted by a solid load that is not anomalous in Hg. The shifts between normal flow and flood events  
368 enhance the erosion of Hg-rich old (syn-mining) terraces, representing the actual overbanks in some part of  
369 the river, causing an alternance of low and high Hg contents along the riverbed. As described in Figure 5 and  
370 pointed out by Pattelli et al. (2014) for the 2012 flood, Hg pulses and floods are almost systematically in phase.  
371 The variability of metal dispersal associated to the effects of flood-sediments sorting and the mixing of  
372 particulate-associated pollutants, may result in changes of one to two orders of magnitude in metal content  
373 over distances of centimetres (Ciszewski and Grygar 2016). Therefore, overbank deposits and channel bars  
374 in the Paglia River represent a secondary source of Hg pollution, leading to the periodical transport of  
375 temporarily stored Hg-rich sediments to the river channel and to the floodplain. This phenomenon prevents a  
376 decrease of Hg concentration over time at least in short time (i.e., decades). Overbank sediments may indeed  
377 represent long-term storage for fine sediments with a residence time of the order of 10<sup>2</sup>–10<sup>3</sup> years (Grygar et  
378 al. 2016). The constant re-mobilization of contaminated material makes the Paglia River system not very  
379 resilient. A similar process is occurring in the Siele Creek, one of the largest Paglia River tributaries (Fornasaro  
380 et al. 2022).  
381 Since the contaminated area along the Paglia River almost corresponds with the area identified by the  
382 hydraulic hazard map of the Tiber River management basin plan (Trigile et al. 2018; Fig. 6), in the next future  
383 it is expected that further Hg mobilization will take place during flood events. The recent broadening of the  
384 Paglia River, started in 2010, coincided with an increase in monthly water discharge observed from 2003,  
385 consequent of a higher frequency of extreme flood events (Pattelli et al. 2014; Rimondi et al. 2014; Cencetti et  
386 al. 2017). These events will be predictably influenced by the variations of the precipitation regime because of  
387 climate change (van Vliet et al. 2013; Papalexiou and Montanari 2019). More precisely in southern Europe and

388 in the Mediterranean region it is expected an overall drastic reduction in precipitation, more pronounced in  
389 summer (-25-30%; Castellari et al. 2014) Regional-scale model projections for Italy show indeed a significant  
390 temperature increases for the period 2070-2100 and a reduction in the number of days with little rain, and, by  
391 contrast, an increase of days with heavy rainfall (Castellari et al. 2014). Frequent drought periods characterized  
392 by long periods of low water flow, with modest or almost no solid transport, will alternate with intense rainy  
393 periods or flash floods, concentrating solid transport in few short events. ~~The contraction of the periods of snow~~  
394 ~~will also affect river dynamics (Billi and Fazzini 2017).~~ Consequently, climate variability could contribute to  
395 control the Hg distribution and overall mobility from MAMD and the Paglia River floodplain up to the  
396 Mediterranean Sea by the way of the Tiber River.

397 Our study provides useful information for management authorities to define precaution actions (such as  
398 limitations of sediment remobilization, river dredging, instream mining) and to identify conservation measures  
399 in this area (e.g., tree planting on overbanks, retention basins, thresholds and/or selective weirs). Further  
400 monitoring is necessary to ensure that the environmental quality of the river will not be altered by the spatial  
401 variability of Hg contaminated sediments distribution. The same strategies can be applied to similar rivers  
402 draining metal-contaminated areas that changed their morphology from braided to narrower channel, which in  
403 time are likely to act as continuous sources of contaminated particles deposited in their abandoned floodplains.  
404 On the other hand, the knowledge of distribution patterns of contaminated sediments is useful to address  
405 geomorphologic issues, as they can represent a tracer within the sediment system, providing a useful marker  
406 to the extent of sedimentation in a certain period. Furthermore, by tracking the movements, re-working, and  
407 removal of these contaminated sediments the role of floodplains as sediment storages can be established at  
408 different timescales.

409

## 410 5. Conclusions

411 The geomorphological and morphodynamic changes of the Paglia River, combined with anthropogenic  
412 activities occurred in the last century, controlled the spatial variability of Hg concentration in channel sediments  
413 and floodplain deposits of the northern stretch of the Paglia River, downstream the Monte Amiata Mining  
414 District. The distribution of Hg observed in the Paglia River floodplain resulted from the interplay of Hg mining,  
415 that fed the floodplain with large amounts of Hg-contaminated sediments during the braided stage (end of  
416 1800-mid-1950s) and the subsequent morphological changes of the river, following World War II (including  
417 gravel mining and other anthropogenic modifications), that led to the single-channel morphology of the Paglia  
418 River. After mine closure, a reduction of Hg concentration over time in river sediments did not occur, as it could  
419 be expected. Because of the braided narrowing morphology, the Paglia River enhanced the erosion of old syn-  
420 mining terraces, rich in Hg, and redistributed Hg contaminated sediments. Consequently, the process of  
421 transport/deposition did not allow a natural “clean up” of the river system since the closure of the mining sites.  
422 The temporal and spatial variability of Hg distribution is therefore principally associated with the fluvial  
423 geomorphological changes more than to anthropogenic activities.

424 At present, the main factor controlling Hg distribution in the next future is identified in climate variability,  
425 triggering erosion/deposition and redistribution of previously stored Hg contaminated ~~sediments in the~~  
426 ~~overbanks overbank sediments~~ and in the floodplain. In the Paglia River upper section, the alternation of

427 normal flow conditions and flood events affects the geomorphology of the river course contributing to make  
428 overbank erosion a permanent secondary source of Hg. The expected intensification of extreme weather  
429 events (high rain events, intense floods), consequent of climate change, makes this area a Hg source of  
430 remarkable environmental concern at the local (Paglia River), regional (Tiber River), and Mediterranean scales  
431 in the future.

432

### 433 **Acknowledgments**

434 We thank Mario Paolieri (Università di Firenze) and members of the ARPAT staff for support in the field and in  
435 the laboratory. We acknowledge the useful comments by the Editor and two anonymous reviewers.

436

### 437 **Funding and Conflicts of interests**

438 The research was funded by a specific agreement between ARPAT (responsible: C.F.) and Università degli  
439 Studi di Firenze (responsible: P.C.). The authors declare that they have no known competing financial interests  
440 or personal relationships that could have appeared to influence the work reported in this paper.

441

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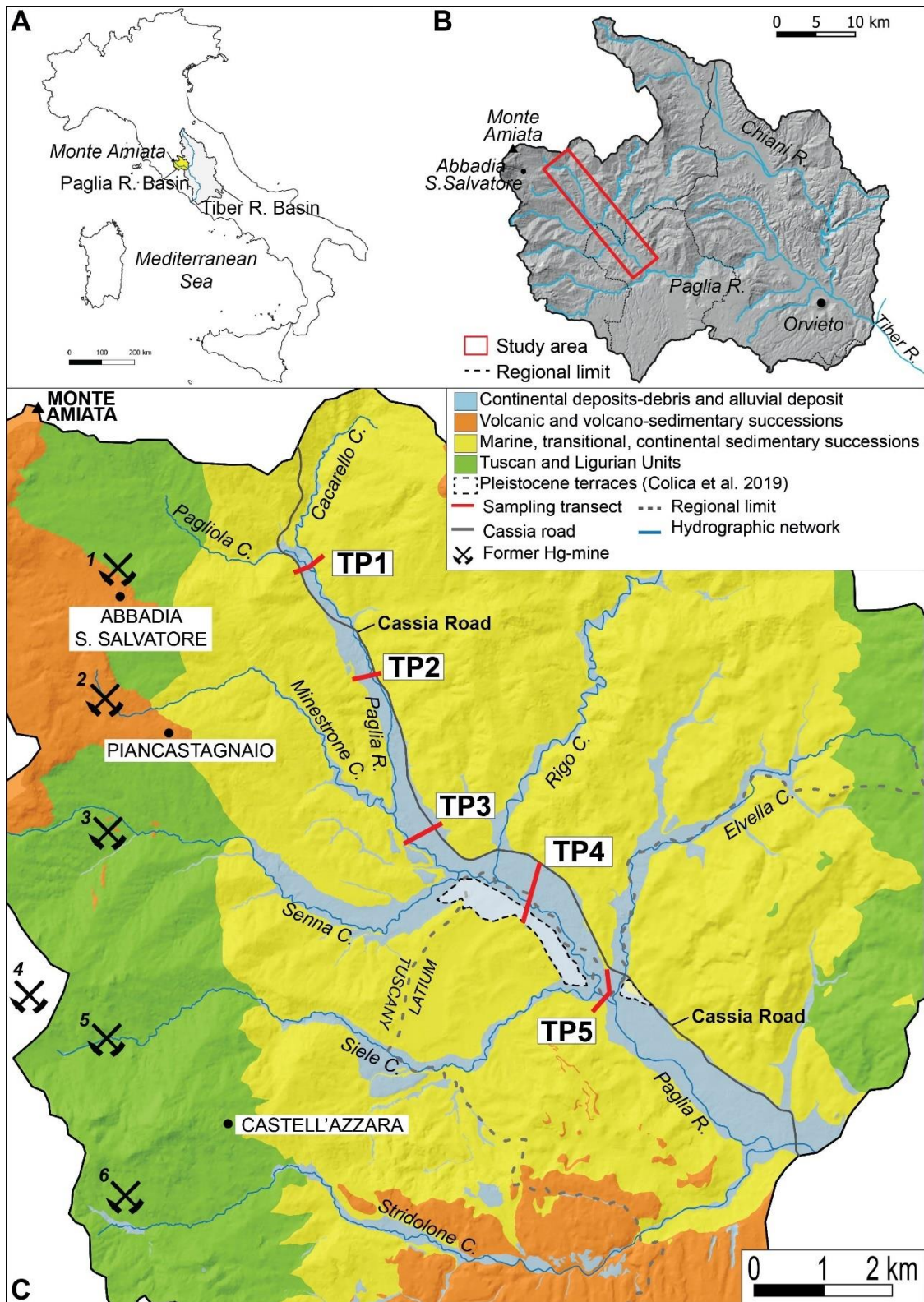
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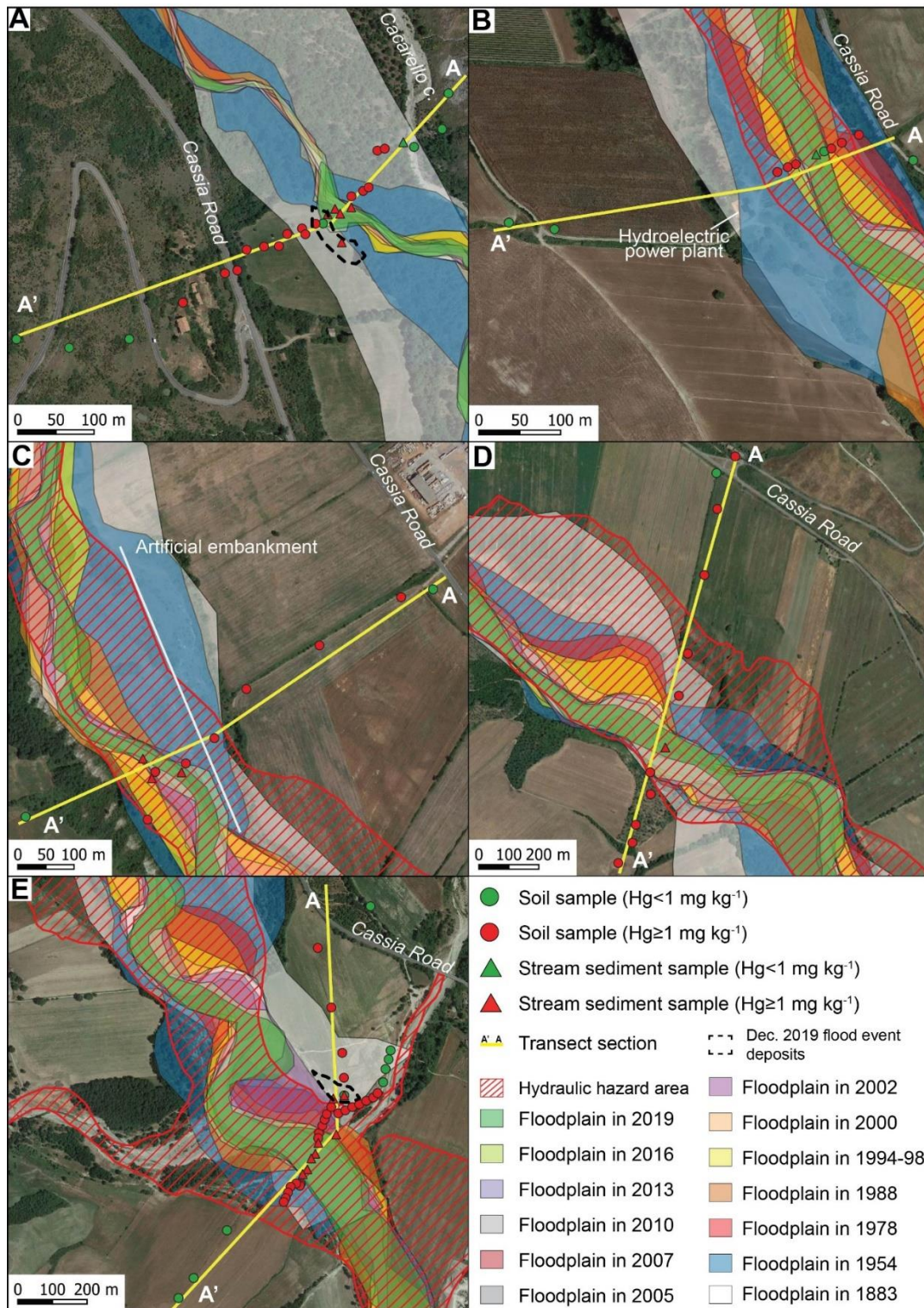
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<b>Element</b>	<b>Definition</b>	<b>Reference</b>
<b><i>Floodplain</i></b>	The floodplain is formed by past active channel riverbed abandonments. Two mechanisms, lateral migration by the braid-train and reactivation of abandoned channels within floodplains, operating separately or in combination, are responsible for floodplain reworking and their relatively young age (<250 years). Clearly, braided rivers can construct substantial areas of well-developed floodplain.	Aute et al. 2010
<b><i>River channel</i></b>	The active channel (or riverbed). The channel through which the water flows.	-
<b><i>Braided channel</i></b>	A network of channels formed in a river that has a great amount of sediment and a fluctuating pattern of discharge: the braiding effect is created by the formation of braid bars, around which the individual channels flow.	-
<b><i>Single channel with low sinuosity</i></b>	Sinuosity defines the degree of meandering of a riverbed. Channel sinuosity arises from flow hydraulic processes around bends in which secondary, across-channel circulation can increase meander wavelength and the migration of meanders across a floodplain. In general sinuosity is low in confined mountain streams.	Leopold et al. 1964
<b><i>Overbank sediment</i></b>	Overbank sediments occur along rivers and streams with variable water discharge. They are deposited on floodplains and levees from water suspension during floods, when the discharge exceeds the amounts that can be contained within the normal channel.	Bolviken et al. 2004
<b><i>Riverbank</i></b>	The landform distinguished by the topographic gradient from the bed of a channel along the lateral land-water margin up to the highest stage of flow or up to the topographic edge, where water begins to spread laterally over the floodplain surface.	Florsheim et al. 2008
<b><i>Bank erosion</i></b>	Bank erosion refers to the erosion of sediment from riverbank.	Florsheim et al. 2008

Tab. 1 Definition of the geomorphological terms used in the text.

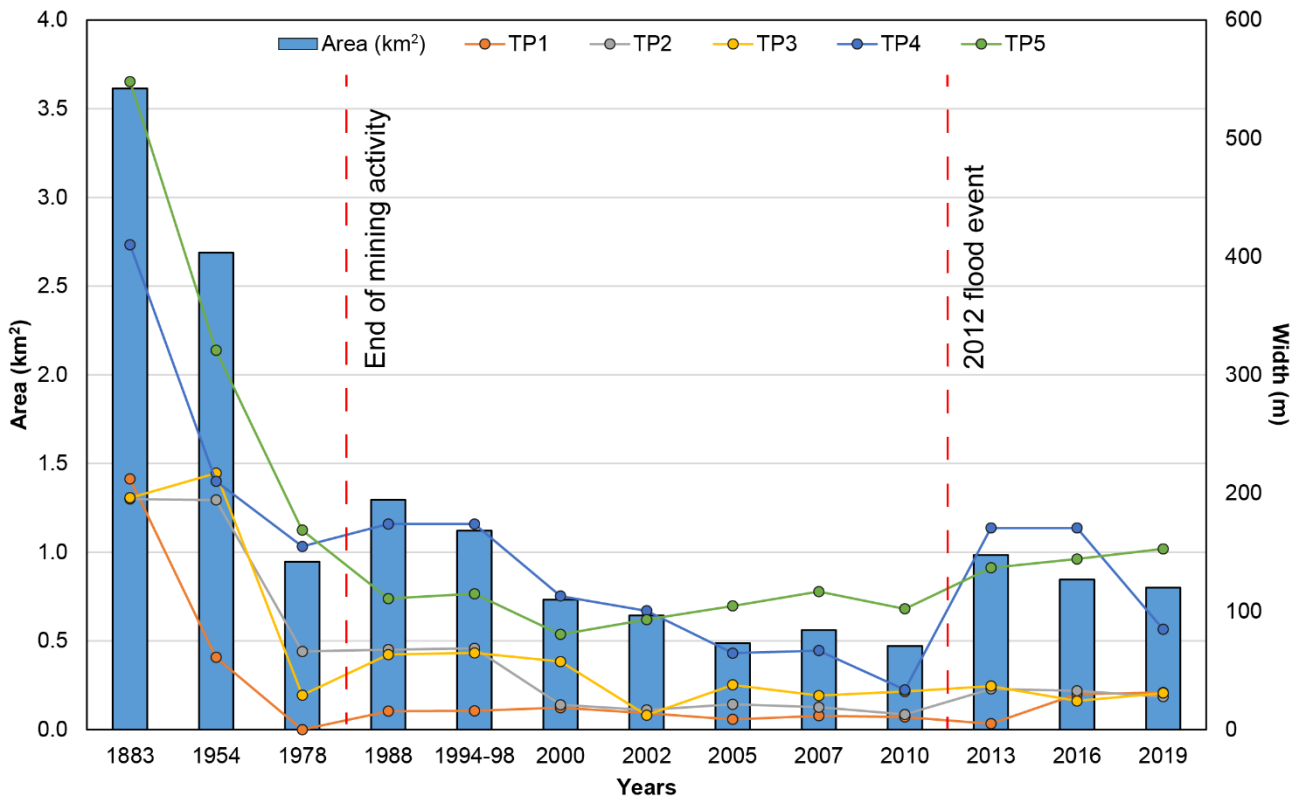


664  
 665 Fig. 1 A) Monte Amiata and Paglia-Tiber system location; B) Paglia River basin and its main tributaries; C)  
 666 Geological map of the upper part of the Paglia River basin. The location of the sampling transects and of the  
 667 main mines of MAMD are also reported: 1) Abbadia S. Salvatore; 2) Case di Paolo - Cerro della Tasca; 3)  
 668 Senna; 4) Solforate; 5) Siele; 6) Cornacchino.



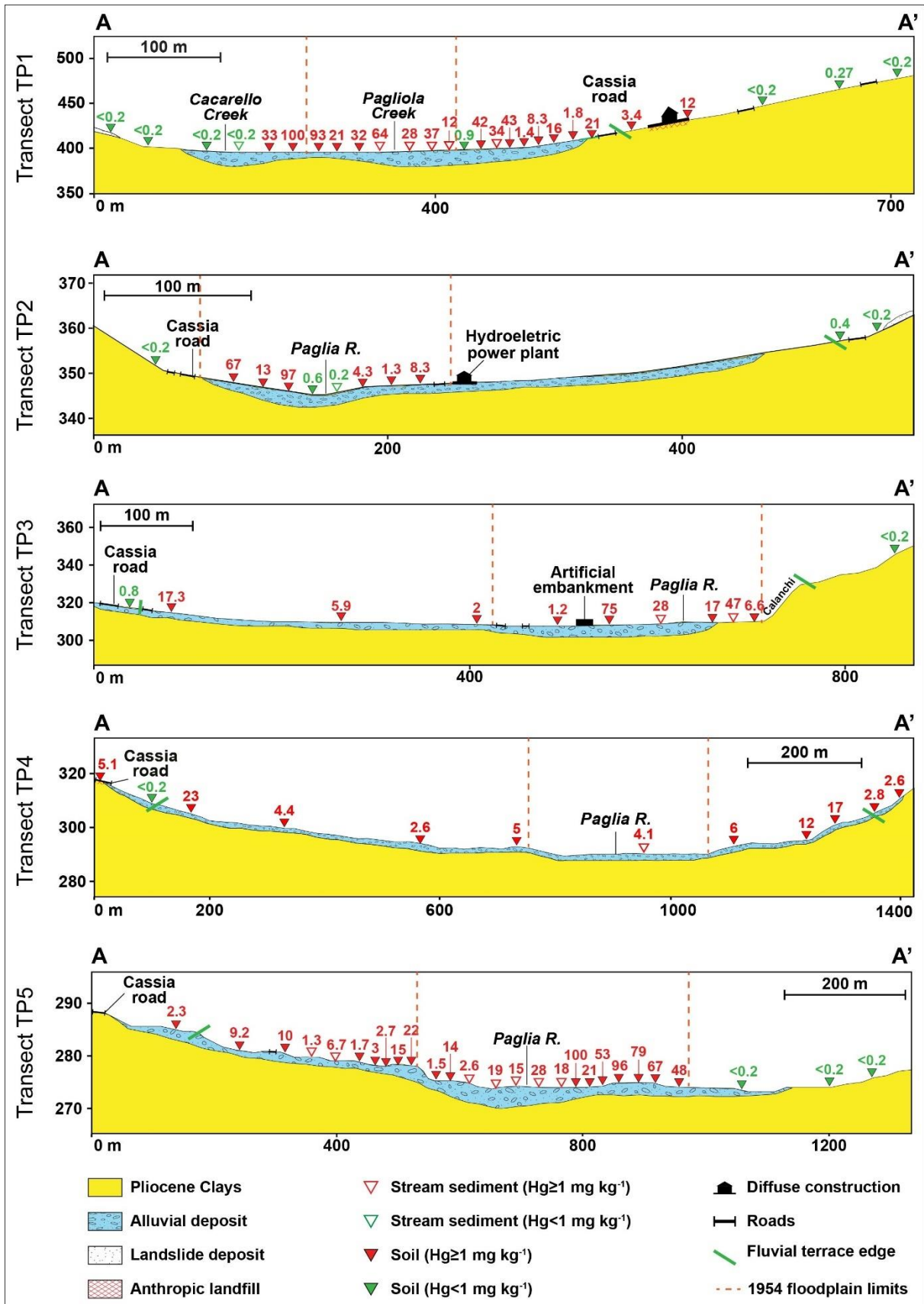
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670 Fig. 2 Satellite image (from Google Maps®, 2019) showing the samples collected along the transects in the  
 671 Paglia R. A) transect TP1; B) transect TP2; C) transect TP3; D) transect TP4; E) transect TP5. Samples with  
 672 Hg concentrations lower and higher than  $1 \text{ mg/kg}^{-1}$  (D.Lgs. 152/2006) are shown in green and red, respectively.  
 673 The boundary of the riverbed (colored area) in different years from 1954 to 2019, the hydrologic hazard area  
 674 (lined area), the transects (yellow lines), and the December 2019 flood event deposits (blue dotted line in A  
 675 and E) are also shown.



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677 Fig. 3 Floodplain area (m<sup>2</sup>) and width (m) along the sampled transects (colored lines) during the 1883-2019  
 678 period.



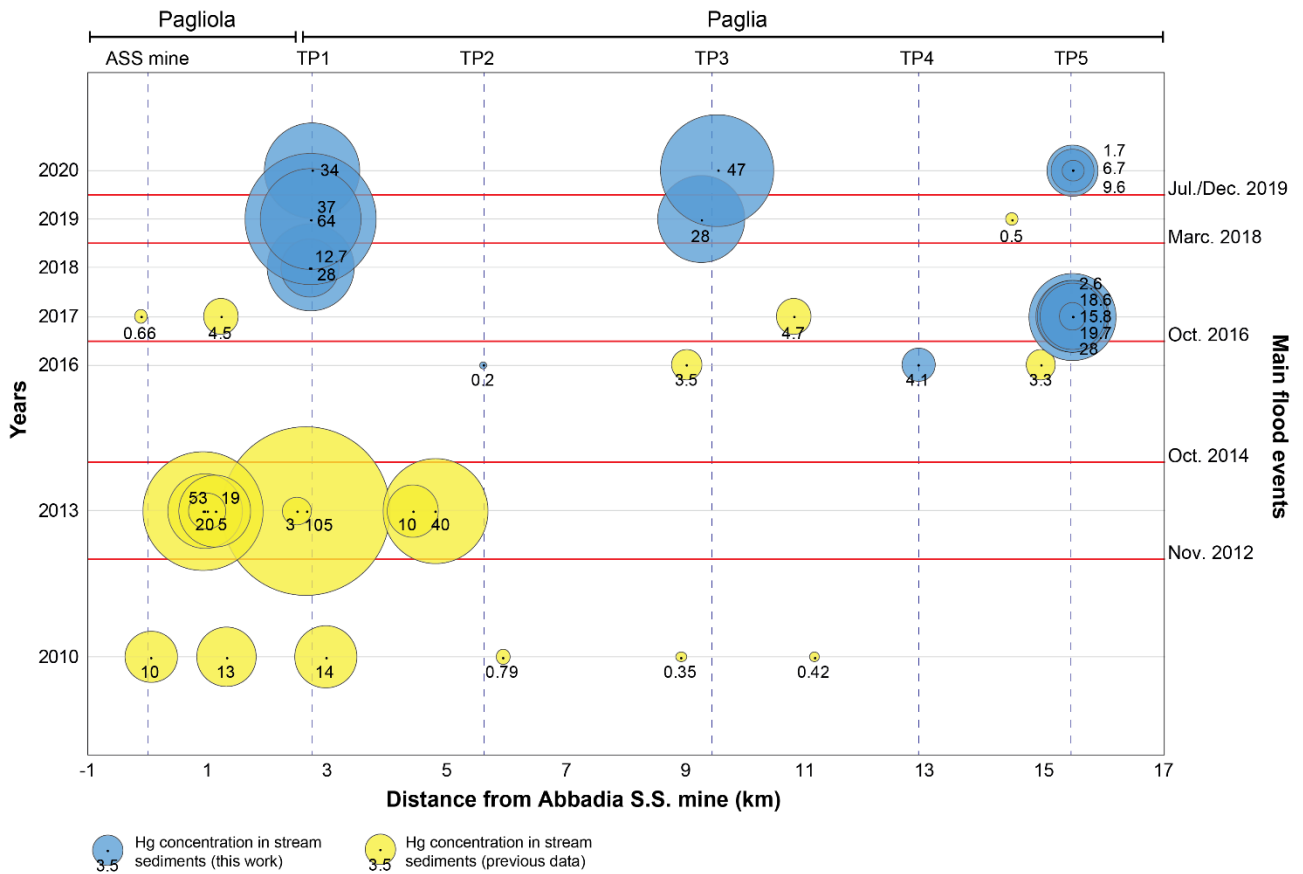
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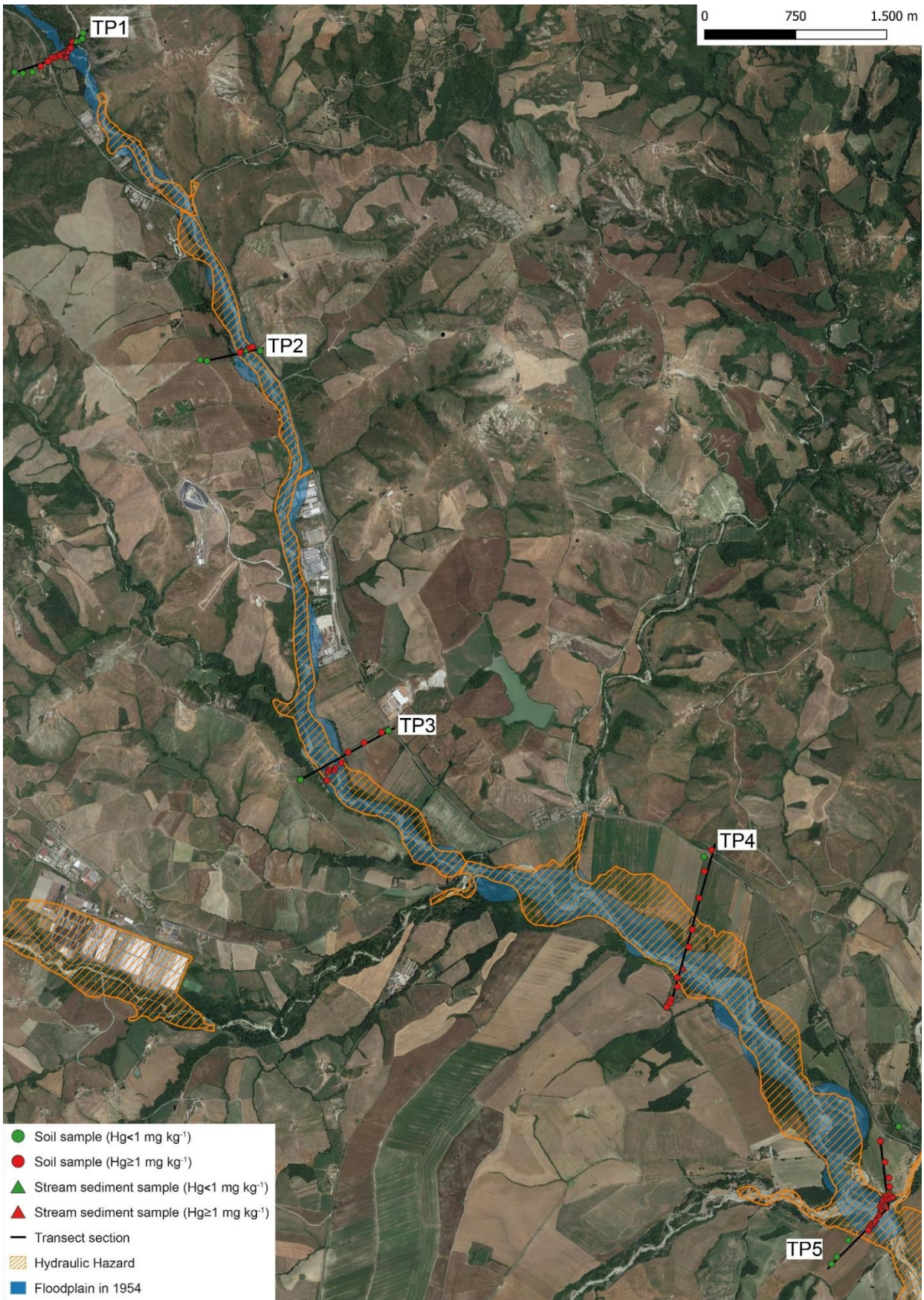
Fig. 4 Geological sections, sample location, and Hg-concentration ( $\text{mg}/\text{kg}^{-1}$ ) in soils and sediments sampled along the transects. View from North to South.





682

683 Fig. 5 Time-space variability of Hg concentrations (mg/kg<sup>-1</sup>) in stream sediments along the Paglia River course  
 684 in different years. Sampling location is indicated on the top X axis. The main flood events are also reported.



686 Fig. 6 Hydraulic hazard area (limits of Triglia et al., 2018) and Hg-concentration in stream and soil samples  
687 ( $\text{mg kg}^{-1}$ ).

