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87	Abstract	<p>Biochar properties are highly dependent on feedstock type and operational conditions during thermochemical processing, in particular slow pyrolysis. To clarify this aspect, nine biochars were produced by pyrolyzing in a macro TGA at 400, 550, and 650 °C three different decorticated and chopped biomasses. The studied biomasses are representative of conifer (black pine) and deciduous (poplar and willow) woods. Biochar surface area, size, and shape of pores were investigated by means of nitrogen adsorption isotherm, Hg porosimetry, and electron microscopy. The results indicate that biochars with high surface area can be obtained at high temperature, especially starting from pine feedstock. Regarding porosity, micro-pores (1–10 nm) are not remarkably affected by the starting feedstocks, while macro-pores (> 10 nm) are strictly connected with the morphology of the starting wood. More than the surface area, we found a strong correlation between the chemical composition (elemental composition and FTIR) of the biochars and their retention and release capacity of ions (cation exchange capacity, CEC). The trend in the CEC, determined via coupled approach by spectrophotometric and ion chromatography, reveals that the increase in the processing temperature has the effect of reducing the number of functional groups able of exchanging the cations with the equilibrium solution. This work represents a step forward in the characterization of the char produced by pyrolysis of biomass thanks to the</p>	

development of a multi-technique approach allowing to obtain a structure-property correlation of the biochars. Our results and experimental approach can help in the optimization of the parameters used in the preparation of these materials.

Graphical abstract: Figure – Left: Correlation of CEC and elemental analysis obtained from willow biochars produced at different pyrolysis temperatures. Right: Scanning electron micrographs at the same magnification (1 kX) of willow derived biochar at different pyrolysis temperatures (panel 1: 400 °C, panel 2: 550 °C and panel 3: 650 °C).s.

88	Keywords separated by ' - '	Biochar - Lab-scale pyrolysis - Porosity - Cation exchange capacity (CEC) - Lignocellulosic biomass - Pyrolysis temperature - Water retention
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89	Foot note information	Giovanni Ferraro and Giuditta Pecori contributed equally to this work.
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Supplementary Information

Supplementary material

The following are the Supplementary data to this article:
Supplementary data: adsorption isotherms related to black pine, poplar, and willow biochar produced at different pyrolysis temperatures: 400 °C, 550 °C and 650 °C. Feedstocks humidity and biochar yields for the different samples: black pine, poplar and willow. Elemental analysis, H/C ratio and C/N ratio of biochars obtained at different temperatures and starting feedstocks. Ash and volatile content of the feedstocks and biochars for the different samples. Amount of readily soluble cations removed in the pretreatment and exchangeable cations obtained after washing with ammonium acetate. (DOCX 241 kb)



Biochar from lab-scale pyrolysis: influence of feedstock and operational temperature

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Abstract

Biochar properties are highly dependent on feedstock type and operational conditions during thermochemical processing, in particular slow pyrolysis. To clarify this aspect, nine biochars were produced by pyrolyzing in a macro TGA at 400, 550, and 650 °C three different decorticated and chopped biomasses. The studied biomasses are representative of conifer (black pine) and deciduous (poplar and willow) woods. Biochar surface area, size, and shape of pores were investigated by means of nitrogen adsorption isotherm, Hg porosimetry, and electron microscopy. The results indicate that biochars with high surface area can be obtained at high temperature, especially starting from pine feedstock. Regarding porosity, micro-pores (1–10 nm) are not remarkably affected by the starting feedstocks, while macro-pores (> 10 nm) are strictly connected with the morphology of the starting wood. More than the surface area, we found a strong correlation between the chemical composition (elemental composition and FTIR) of the biochars and their retention and release capacity of ions (cation exchange capacity, CEC). The trend in the CEC, determined via coupled approach by spectrophotometric and ion chromatography, reveals that the increase in the processing temperature has the effect of reducing the number of functional groups able of exchanging the cations with the equilibrium solution. This work represents a step forward in the characterization of the char produced by pyrolysis of biomass thanks to the development of a multi-technique approach allowing to obtain a structure-property correlation of the biochars. Our results and experimental approach can help in the optimization of the parameters used in the preparation of these materials.

Keywords Biochar · Lab-scale pyrolysis · Porosity · Cation exchange capacity (CEC) · Lignocellulosic biomass · Pyrolysis temperature · Water retention

1 Introduction

Biochar is the carbon-rich product obtained by heating lignocellulosic biomass in the absence of (or with very limited) oxygen at process temperatures normally in the range of 400–600 °C [1]. The word “biochar” is a neologism that combines the words “bio” (from Greek, life) and “char” (from English, charcoal, to distinguish from fossil coal). The definition was proposed by IBI (International Biochar Initiative) and specifically indicates that this material is used in the field of agricultural and environmental protection [2], and then it was extended to many other applications far beyond agriculture, as flue gas and water treatment (most of these uses are well known and fully mature at industrial scale since decades). Currently, several thermochemical technologies such as pyrolysis, gasification, and hydrothermal conversion are employed to produce biochar. Various types of biomass can be used as feedstock, including agricultural and forestry

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48 products and by-products, such as wood chips, straw, nut
49 shells, rice hulls, tree bark, wood pellets, and switch grass;
50 agro-industrial and biorefining by-products, as sugarcane ba-
51 gasse and straw, paper sludge, and cellulose pulp; animal
52 wastes, such as chicken litter, dairy and swine manure; and
53 sewage sludge. Producing biochar from lignocellulosic bio-
54 mass, especially forest and agricultural wastes, represents an
55 excellent way to valorize residues into bio-based materials [2].

56 Biochar is composed of carbon (C), hydrogen (H), oxygen
57 (O), nitrogen (N), phosphorus (P), sulfur (S), and ash in dif-
58 ferent proportions. Its properties extensively change depend-
59 ing on the type of feedstock (wood structure, chemical com-
60 position, ash content, particle size), the process conditions
61 (temperature, time, oxidative conditions), the pre-treatment
62 (drying, crushing), and post-treatment steps (for instance, ac-
63 tivation methods if the final product is activated carbon) [3].
64 The presence of pores of different diameter ranging from less
65 than 2 nm up to more than 10 μm [1] in biochar provides high
66 adsorptive capacities to this material and allows the adsorption
67 of small molecules, such as gases and solvents [1]. Thus,
68 biochar is an effective material for different applications, in-
69 cluding waste management, soil remediation, and carbon se-
70 questration [4–7]. Furthermore, biochar has been reported to
71 improve soil fertility and quality by raising soil pH, and in-
72 creasing moisture holding capacity and plant-available water
73 [8, 9]. The surface chemistry of biochar (e.g., the ratio
74 hydrophilic/hydrophobic domains) is another parameter that
75 could affect the ability to retain and release water [1]. Suliman
76 et al. demonstrated that the increase in the oxygen functional
77 groups on the surface of biochar enhances the water-holding
78 capacity of the material [3]. Finally, it is proved that the use of
79 biochar as soil amendment significantly decreases soil bulk
80 density, promotes the soil organic matter [10], attracts more
81 beneficial fungi and microbes [11], and retains nutrients (po-
82 tassium, phosphorus, zinc, calcium, and copper) [12, 13]. It is
83 therefore of great interest to study biochar's ability in cations
84 retention and release: cation exchange capacity, cation ex-
85 change capacity (CEC), is defined as the total amount of ex-
86 changeable cations that the soil can adsorb. In the soil, cations
87 such as Ca^{2+} , Mg^{2+} , K^+ , Na^+ , H^+ , Al^{3+} , Fe^{3+} , and Mn^{2+}
88 are retained both by the clay, negatively charged for the presence
89 of hydroxyl groups present in the phyllosilicates, and organic
90 matter through electrostatic forces. In the case of organic mat-
91 ter, the cation exchange capacity is mainly due to the presence
92 of carboxyl groups. The cations in the soil are easily ex-
93 changeable with the equilibrium aqueous phase and can be
94 adsorbed by the roots of the plants [14].

95 Several methods have been proposed for the determination
96 of the cation exchange capacity of biochar, depending on the
97 particular feedstock and production technique employed
98 [15–17]. In addition, potential sources of error in CEC deter-
99 mination could arise from the biochar microporous structure,
100 which can prolong the equilibration time, and from its intrinsic

hydrophobicity, which can cause poor wetting of the sample
with a global underestimation of the CEC value. The presence
of base cations, such as those linked to carbonates and sili-
cates, can interfere with the sum of exchangeable base cations
giving an overestimation of CEC [18]. For these reasons,
Munera-Echeverri et al. [19] tried to modify and critically
assess different steps in the ammonium acetate (NH_4OAc)
method (pH 7). They introduced a pretreatment step of bio-
char, using diluted hydrochloric acid, to decrease biochar pH
to near neutral, so that 1 M NH_4OAc effectively buffers the
biochar suspension pH at 7. This allows the CEC of all bio-
chars to be determined at pH 7, which is crucial for biochar
comparison. Skipping the pretreatment step causes a major
overestimation of the CEC of biochar. Moreover, they ob-
served that isopropanol may not penetrate the smallest pores
of some biochars and therefore other liquid compounds are
required to fully remove excess NH_4OAc .

In this study, the macro- and micro-porosity of nine bio-
chars, produced by pyrolysis at different temperatures
(400 °C, 550 °C, and 650 °C) of both softwood (black pine)
and hardwood (poplar and willow) were investigated. The
combined use of nitrogen adsorption isotherm, Hg
porosimetry, and electron microscopy was proposed to inves-
tigate the relationships between the feedstock, the pyrolysis
temperature, and the physico-chemical properties of the ob-
tained char. Furthermore, the retention and release capacity of
ions was quantified via CEC using the ammonium acetate
method, as usually determined in soil analysis and adapted
for biochar. The final goal was to contribute into the under-
standing of the correlation between the chemical composition
and micro-/macro-structure of the char with its retention/
release capacity.

To the best of our knowledge, this work represents a step
forward in the biochar field since previous works report only a
partial picture of the problem. Indeed, some articles investi-
gated the properties of biochar produced from various feed-
stocks at different pyrolysis temperatures in terms of biochar's
water retention [20] or connected to its mechanical properties
[21]. Other authors focused their attention only on the CEC
obtained on different biochars [15, 18, 19].

2 Material and methods

2.1 Biochar production

Biochar samples were produced from woody feedstocks:
black pine, poplar, and willow woods ~ 15 cm length and 1–
2 cm diameter were peeled, milled, and sieved at dimension of
4 mm using a Retsch SM 300 mill. The biomass was not
pretreated before pyrolysis. Pyrolysis was performed in a mac-
ro TGA (LECO TGA 701) under nitrogen flow (10 L/min) at
a heating rate of 20 °C/min, maintaining a 2-h thermal plateau

150 at 400 °C, 550 °C, or 650 °C. All biochars were prepared
 151 starting from 2 ± 0.1 g of feedstock and using a large TGA
 152 ceramic crucible (volume 20 mL). All samples were
 153 characterized in terms of elemental analysis, surface area,
 154 pore size distribution, and functional groups and
 155 tested to determine the retention and release capacity
 156 of ions via CEC. The appearance of the samples at
 157 the different production steps is reported in Fig. 1.

158 The samples will be referred as black pine (BP), poplar (P),
 159 and willow (W).

160 **2.2 Chemicals**

161 Isopropanol (98%), potassium chloride (KCl, 99%), ammonium
 162 acetate (NH₄OAc, 98%), hydrochloric acid (HCl, 37%),
 163 sulfuric acid (H₂SO₄, 95–97%), sodium nitroprusside (99%),
 164 and sodium salicylate (99.5%) were purchased from Sigma-
 165 Aldrich and used as received without further purification.

166 **2.3 Elemental analysis (CHNS-O-ashes)**

167 Elemental analysis was performed using the instrument
 168 LECO TruSpec CHN equipped with the TruSpec S modulus
 169 on 60–80 mg of crushed samples by a routine flash combustion
 170 procedure. Each sample was measured three times. The
 171 instrument was calibrated with a phenylalanine standard for
 172 the CHN modulus while coal was used as standard for the S

173 modulus. The oxygen (O) content was inferred by difference
 174 as $O = 100 - (C + N + H + S + \text{ash})$. The ash content was measured
 175 using thermogravimetric analysis (LECO TGA 701)
 176 after moisture removal (105 °C under air flux 5 L/min) and
 177 oxidative heating at 550 °C (7 L/min oxygen flow) according
 178 to the procedure UNI EN ISO 18122. Volatile components
 179 were quantified heating the samples at 900 °C under nitrogen
 180 flow (10 L/min) for 7 min (procedure UNI EN ISO 18123).

181 **2.4 Surface area and pore size distribution**

182 Surface area was determined by N₂ adsorption isotherms
 183 (BET) in a Quantachrome NOVA 2200E instrument.
 184 Experiments were performed on 60 mg samples preliminarily
 185 dried at 200 °C for 48 h. All measurements were performed
 186 after degassing (200 °C for 24 h). Micro-porosity was estimated
 187 using the DFT approach on BET isotherms as reported
 188 elsewhere [22], while the macro-porosity of biochar samples
 189 produced at 550 °C was also evaluated using a PoreMaster-60
 190 Hg porosimeter.

191 **2.5 Scanning electron microscopy**

192 SEM experiment was conducted using a ΣIGMA high-
 193 resolution scanning electron microscope (Carl Zeiss) based
 194 on the GEMINI column which features a high-brightness
 195 Schottky field emission source, beam booster, and in-lens

Fig. 1 Black pine (panel a),
 poplar (panel b), and willow
 (panel c) woods before (top) and
 after (middle) the milling process.
 Biochars obtained after the
 pyrolysis are shown in the bottom
 line



196 secondary electron detector. Measurements were conducted
197 on uncoated samples with an acceleration potential of 2 kV
198 and at a working distance of about 3 mm.

199 2.6 Infrared spectroscopy

200 Infrared spectroscopy was performed by a SHIMADZU
201 IRTracer-100 spectrometer. Data were collected at room tem-
202 perature in attenuated total reflectance (ATR) mode. The op-
203 tical resolution was 4 cm^{-1} , and the spectral range investigated
204 was from $600\text{ to }4000\text{ cm}^{-1}$. A total of 45 scans was averaged
205 to have an acceptable signal-to-noise ratio.

206 2.7 Ion chromatography

207 The ion chromatograph used for the analysis was a Dionex
208 DX120 equipped with a Dionex Ion Pac CG12A ($4 \times 50\text{ mm}$)
209 guard column and a Dionex Ion Pac CS12A ($4 \times 250\text{ mm}$) sep-
210 aration column (eluent: H_2SO_4 22.5mN). This experimental set-
211 up allows to obtain a good detection of ammonium if compared
212 to other cationic species with similar retention times (Na^+ and
213 K^+). However, in order to eliminate the interference of K^+ ions,
214 all samples were diluted 500–1000 times with MilliQ water.

215 2.8 Cation exchange capacity procedure

216 The method used for the determination of the cation exchange
217 capacity was the one proposed by Munera-Echeverri et al.
218 [19]. The procedure can be briefly summarized in three main
219 steps: pretreatment, release of exchangeable cations in
220 NH_4OAc , and substitution of NH_4^+ by K^+ . In the first step,
221 biochar samples were washed with water, and pH adjusted to
222 pH 7 using hydrochloric acid (HCl) 50 mM. Next, biochar
223 was washed three times with deionized water (20 mL of
224 $\text{H}_2\text{O}/1\text{ g}$ of solid) until conductivity values around $0.2\text{ mS}/$
225 cm . In the second step, biochar was extracted two times with
226 NH_4OAc 1 M solution (20 mL/1 g of solid). During the ex-
227 traction, the sample was orbitally shaken for 24 h at 200 rpm
228 and the supernatant was collected after centrifugation at 1700g
229 for 20 min. The supernatant was analyzed for base cations
230 (CEC-BC) using the ICP MP-AES AGILENT 4200. The in-
231 strument was calibrated using a multi-element standard solu-
232 tion for ICP (TraceCERT®, Sigma-Aldrich).

233 In the last step, the excess of NH_4OAc was removed by
234 washing with isopropanol and subsequently the sample was in-
235 cubated with a solution of KCl 2 M (20 mL/1 g of solid) to
236 exchange NH_4^+ with K^+ . The ammonium amount was quantified
237 using two different approaches: ion chromatography (see
238 Section 2.7) and a colorimetric method previously described in
239 the literature [23, 24] with some modifications. Briefly, a reagent
240 A was prepared by dissolving 1.0 g of salicylic acid and 0.1 g of
241 sodium nitroprusside in 100 mL of citrate buffer (0.27 M
242 trisodium citrate and NaOH 0.054 M). Reagent B was prepared

by dilution of 2 mL of 6% sodium hypochlorite (NaOCl) in 243
100 mL of water. A calibration curve was made, using standard 244
solutions of NH_4Cl containing 0, 50, 75, 100, 500, and 1200 245
 NH_4^+ $\mu\text{g}/\text{L}$. Next, 0.5 mL of reagent A and B was added to 246
10-mL plastic tubes containing 3 mL of the KCl 2 M extracts 247
of each sample diluted 1000 times with MilliQ water. The dilu- 248
tion was necessary to obtain a concentration of ammonium ion in 249
the calibration range. The samples were shaken using an orbital 250
shaker, and after 3 h, the absorbance values were read at 655 nm. 251

252 3 Results and discussion

253 3.1 Biochar characterization

254 Table 1 lists the humidity of the feedstocks (calculated following
255 the procedure UNI EN ISO 18134-2) and production yields.

256 As pyrolysis temperature increases, biochar yield decreases
257 for all feedstocks, in agreement with the literature [1, 25] and
258 experience. Instead, the type of feedstock does not have a
259 remarkable influence on biochar yield in this lab-scale TGA
260 experiments. The yields are comparable to those reported for
261 similar systems in the literature [26].

262 Concerning the composition of obtained chars, the results
263 of the elemental analysis carried out on biomasses and bio-
264 chars are shown in Fig. 2 and in Table S1 in ESI file, while the
265 ash and volatile content of the feedstocks and biochars are
266 reported in Table S2 in ESI file. The values of elemental
267 analysis and ash content are given in % mass, dry basis
268 (wt%, db) while the volatile amounts are given in % mass,
269 dry ash free (wt%, daf).

270 The ash content is higher in the case of hardwoods (poplar and
271 willow) compared to softwoods (black pine) both in the starting
272 feedstocks and in the biochars. The amount of ash increases with
273 increasing the pyrolysis temperature, and the willow is the bio-
274 mass which produces more ash followed by poplar and black
275 pine. The opposite trend is found for volatile content of the
276 feedstocks, with black pine containing the highest amount of
277 volatiles. All feedstocks contain about 6 wt% of hydrogen, while
278 the amount of carbon is greater for pine (about 52 wt%) com-
279 pared to poplar and willow with about 49 wt%.

280 As regards biochars, the process temperature influences the
281 elemental composition. Higher temperatures lead to higher
282 devolatilization and thus release of hydrogen and oxygen with
283 a linear trend, while the carbon content shows an opposite
284 trend (in percentage). Thus, the H/C ratio decreases with in-
285 creasing pyrolysis temperature as shown in Fig. 3a. A higher
286 carbon content in the feedstock also leads to a more carbona-
287 ceous biochar (see pine char).

288 The sulfur content is equal to 0.01 wt% for pine and poplar
289 feedstocks and remains constant even in chars produced at
290 different temperatures. Willow contains 0.04 wt% of S,

t1.1 **Table 1** Feedstock humidity and
t1.2 biochar yields for the different
t1.3 samples: black pine (BP), poplar
(P), and willow (W)

Feedstock humidity %			400 °C (yield*)			550 °C (yield*)			650 °C (yield*)		
BP	P	W	BP	P	W	BP	P	W	BP	P	W
28.5	39.4	40.2	29.6	29.9	29.2	19.6	19.7	21.2	17.9	18.7	16.9

*The values of the yields are given in % mass, dry basis (wt%, db)

291 slightly increasing in biochars produced at 400 and 550 °C up
292 to 0.06 wt% as a consequence of the biomass conversion.

293 Figure 3b gives the C/N ratio for all the analyzed samples.
294 This parameter is generally used to predict the exchange and
295 retention of nitrogen in soil: soils with high C/N lead to rapid
296 nitrogen immobilization which in turns avoid nitrogen
297 leaching in the soil and its volatilization [27].

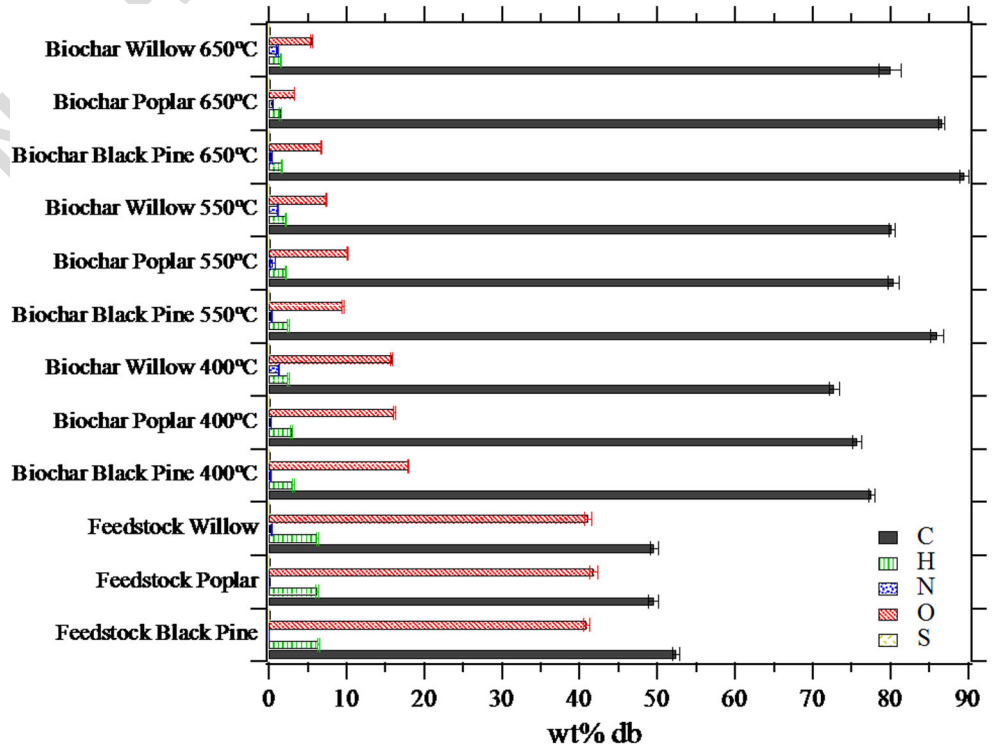
298 However, the C/N ratio of the soil is not the only relevant
299 parameter, as variations of soil temperature and humidity can
300 also affect the final C/N ratio stimulating or inhibiting the
301 microbial activity [28].

302 The main functional groups of the different samples can be
303 evidenced from the IR spectra showed in Fig. 4 panel a while
304 Fig. 4 panel b reports how the surface area obtained through
305 nitrogen adsorption isotherms changes as a function of the
306 production temperature.

307 The spectra of feedstocks obviously differ from those of bio-
308 chars, and the effect of pyrolysis temperature is also evident. The
309 most remarkable change in the char spectra, compared to feed-
310 stocks, is the disappearance of the band at around 3400 cm⁻¹
311 ascribed to the O–H stretching of the hydroxyl groups in

cellulose and hemicellulose and hydration water. The same effect 312
is observed for the signals of hemicellulose and cellulose C–H 313
stretching located at 3000–2800 cm⁻¹. This behavior can be 314
explained as a result of the pyrolysis process [29, 30]. The signal 315
observed at around 1650 cm⁻¹ is clearly visible in the char pro- 316
duced at 400 °C: it is associated to the carboxyl groups of hemi- 317
cellulose, and its intensity decreases as the pyrolysis temperature 318
increases as a consequence of the thermal decomposition of these 319
groups. The evolution of the infrared spectra associated to the 320
gradual removal of oxygen-containing species is in good agree- 321
ment with the findings from the elemental analysis of the chars 322
(Fig. 2). In addition, the samples produced at high temperatures 323
(550 and 650 °C) show the appearance of two new signals in the 324
region 1000–1100 cm⁻¹ and 1300–1500 cm⁻¹ probably ascrib- 325
able to the presence of C–O and C=O groups from ether-like and 326
ketone-like species [31, 32], which are less evident for the char 327
produced starting from black pine feedstock. The decomposition 328
of cellulose, hemicellulose, and lignin is also confirmed by the 329
disappearance of the prominent band centered at around 330
1000 cm⁻¹ in the spectra of the feedstocks, which is assigned 331
to C–O and C–C stretching or C–OH bending [33]. 332

Fig. 2 Elemental analysis of
biochars obtained at different
temperatures and starting
feedstocks



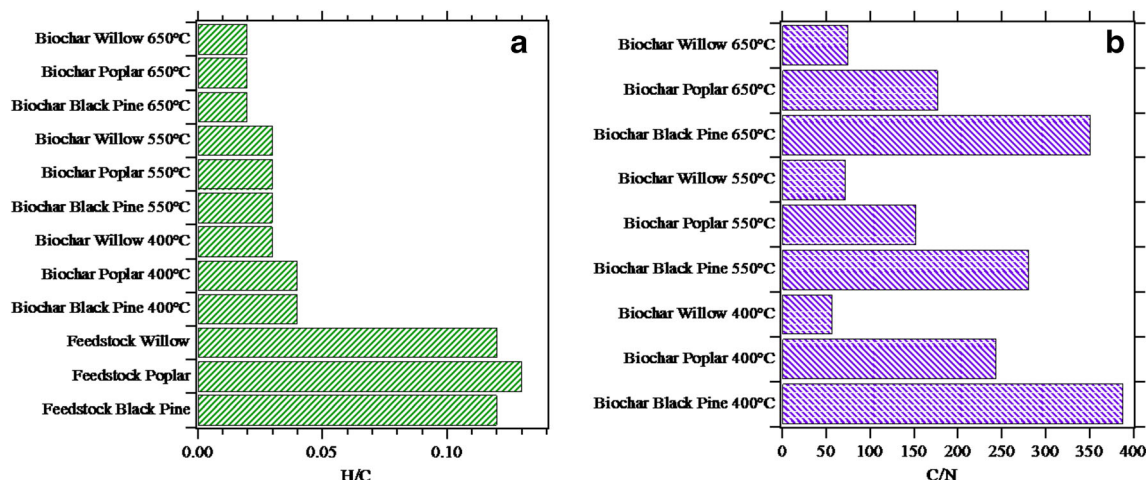


Fig. 3 H/C ratio (panel a) and C/N ratio (panel b) for biochars obtained at different temperatures and feedstocks

333 The bands associated with the aromatic C=C
 334 stretching (1440 cm^{-1}) are clearly visible in biochars'
 335 spectra, and their intensity increases as the working
 336 temperature passes from 400 to 650 °C as a further
 337 confirmation of the graphitization process [34].

338 The surface area and pore volume of chars are among the
 339 most important features affecting the absorption and retention
 340 properties of these materials. Biochars with high surface area
 341 can thus be obtained at higher T (see Fig. 4 panel b). This
 342 result is consistent with the work of McLaughlin et al. [35],
 343 where the increase in the surface area at temperatures above
 344 300 °C can be associated to the development of turbostratic
 345 regions inside the char with a structural organization between
 346 that of amorphous and crystalline carbon, as the solid enriches
 347 in carbon content [1]. However, they demonstrated that above
 348 700 °C the surface area tends to decrease due to the “calcina-
 349 tion” of the graphene residues with the consequent coales-
 350 cence in denser, less porous structures.

351 The effect of high temperatures on surface area is more
 352 evident for pine biochar, which shows the largest BET area
 353 ($504\text{ m}^2/\text{g}$) at 550 °C. This is in good agreement with previous
 354 literature, where it is reported that biochar obtained from

coniferous lignocellulosic biomass is characterized by a great-
 355 er surface area if compared to deciduous equivalents [36]. 356

357 The adsorption isotherms of the biochars at the different
 358 pyrolysis temperatures (see Fig. S1 in the [Supplementary](#)
 359 [Material](#) file) show a hysteresis, indicating the co-presence
 360 of micro- and mesopores in all samples [37]. By using the
 361 DFT approach, the micro-pore size distribution was estimated,
 362 as reported in Fig. 5 panel a, while Fig. 5 panel b reports the
 363 macro-pore distribution obtained from Hg porosimeter on the
 364 biochars produced at 550 °C.

365 The micro-porosity of the chars produced at 400 °C is in
 366 the range between 1 and 10 nm, with no clear differences both
 367 in pore size distribution on the base of the different feedstocks.
 368 By increasing the carbonization temperature, pore sizes in the
 369 range of 1 to 8 nm for black pine were observed, showing the
 370 highest micro-porosity in good accordance with the greater
 371 surface area found for pine biochar.

372 Regarding macro-porosity of biochars produced at 550 °C,
 373 poplar and willow biochars possess pores in the range be-
 374 tween 0.5 and 500 μm with a small population centered at
 375 around 1 μm . Differently, the pore distribution for pine char
 376 shows a pronounced maximum at around 10 μm together with

Fig. 4 Panel a FTIR spectra of black pine, poplar, willow biochars produced at different pyrolysis temperatures (400 °C, 550 °C, and 650 °C), and initial feedstocks. Curves are offset along y-axis. Panel b Surface area from nitrogen absorption measured on black pine (circles), poplar (squares), and willow (triangles) biochars at different pyrolysis temperatures. In the case of 400 °C, error bars are comparable to the marker size

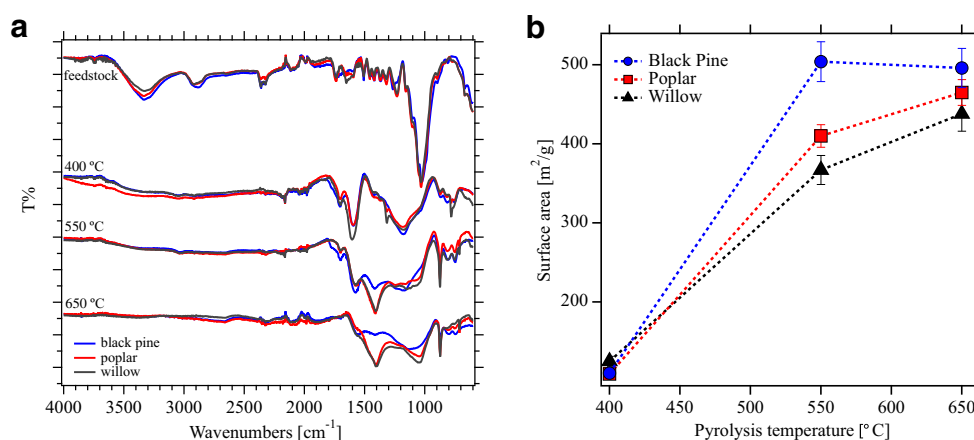
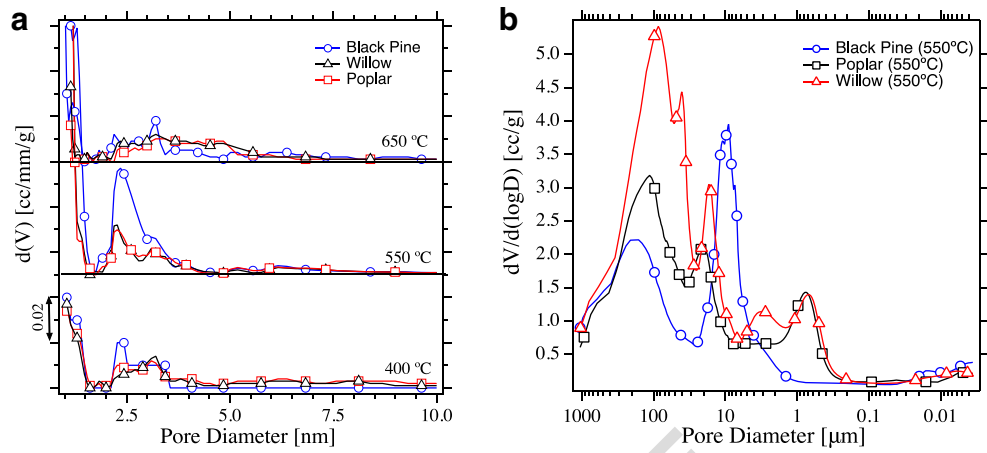


Fig. 5 Micro- (panel a, left) and macro-porosity (panel b, right) obtained from nitrogen and Hg porosimetry, respectively



377 larger pores with a diameter around 200 μm. The observed
 378 differences in the macro-porosity measured for the three chars
 379 suggest a strong dependence of this parameter on the morpho-
 380 logical structure of the feedstock, as already known in the
 381 literature [37]. A further confirmation of the presence of
 382 macropores in the samples comes from scanning electron mi-
 383 crographs reported in Fig. 6.

384 From Hg porosimetry, it is possible to obtain also the
 385 values of density associated to the different samples. In

Fig. 7 the density of the samples versus the surface area ob-
 tained from BET analysis is reported.

Observing the Fig. 8, it is interesting to note that willow
 biochar has the highest macro-porosity (i.e., low density) and
 the lowest surface area. This result might be counterintuitive;
 however, it can be rationalized considering that the surface
 area mainly depends on the micro-porosity, while the density
 is mainly a function of the macro-porosity of the starting bio-
 mass. From the obtained data, we can conclude that willow

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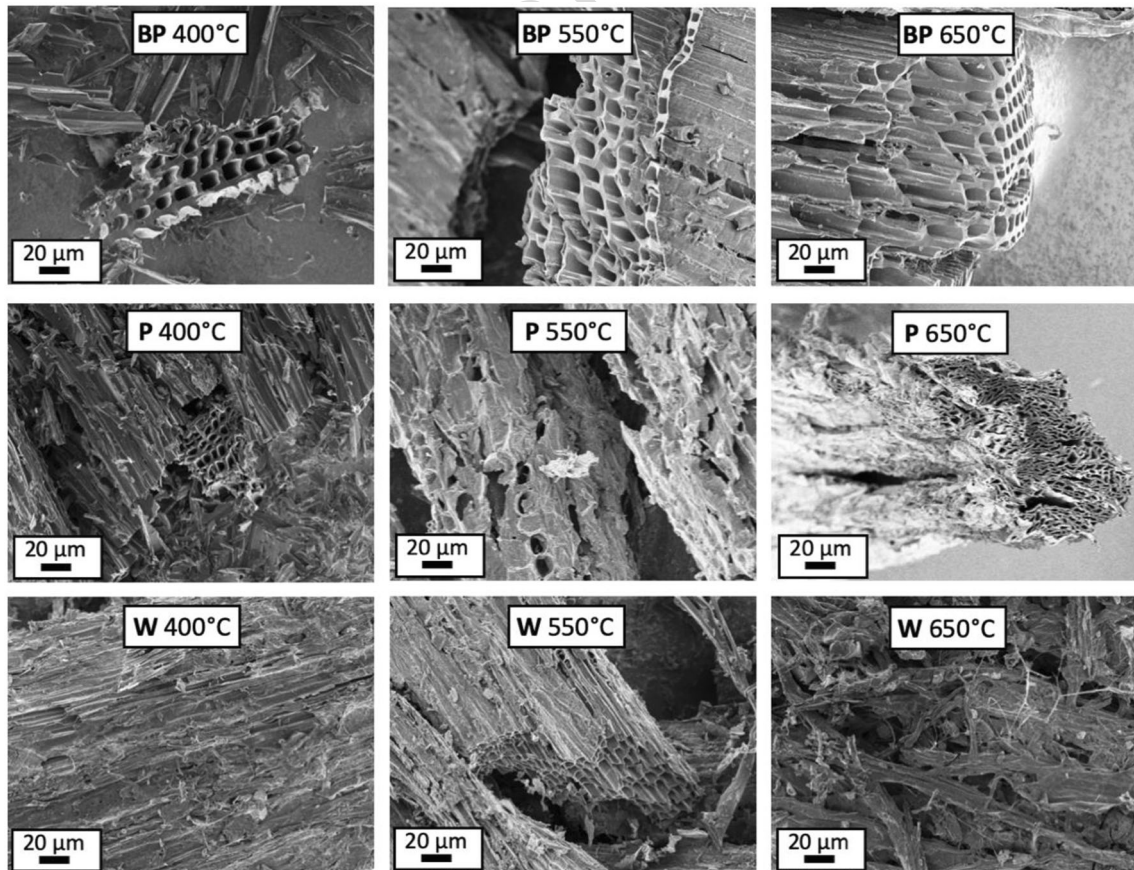


Fig. 6 Scanning electron micrographs at the same magnification (1 kX) of black pine (BP), poplar (P), and willow (W) biochars obtained at different pyrolysis temperatures

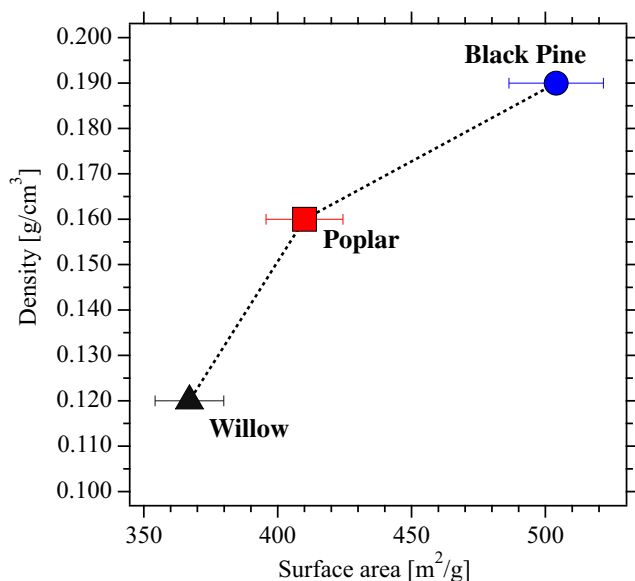


Fig. 7 Density values (from Hg porosimetry) vs surface area (BET) for the char samples prepared at 550 °C

biochar could be the best for retaining water and making it available for the plants. Indeed, it was demonstrated that pores in the nanometric size range are not relevant for water release in the soil because plants are unable to overcome the high capillary forces of water confined in small pores [38].

3.2 Cation exchange capacity

The cation exchange capacity (CEC) was evaluated on the biochars by measuring the base cation exchange capacity (CEC-BC). In the work by Munera-Echeverri et al. [19], remarkable differences between CEC-NH₄⁺ and CEC-BC were

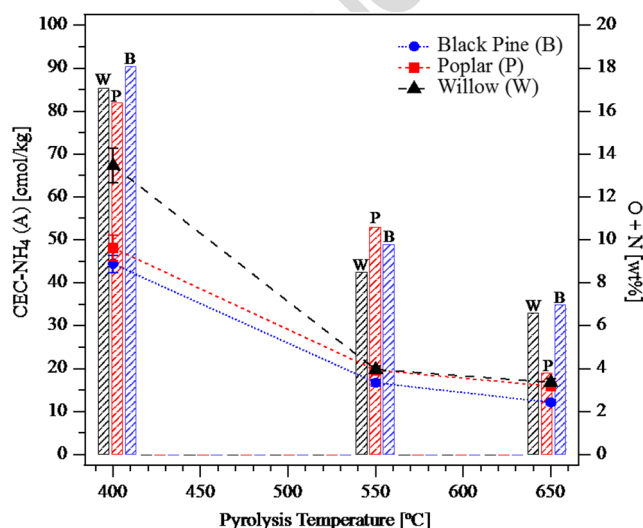


Fig. 8 CEC-NH₄ (markers) and oxygen + nitrogen amounts (bars) obtained from black pine (B, blue), poplar (P, red), and willow (W, black) biochars produced at different pyrolysis temperatures. In the case of 550 and 650 °C, error bars are comparable to the marker size

found, with CEC-NH₄⁺/CEC-BC ratios in the range 1.0–4.0. These discrepancies between the two methods were ascribed to the fact that isopropanol, used as washing agent, does not effectively remove NH₄⁺ trapped in the small pores of biochar, presumably due to the size of isopropanol molecules which cannot penetrate in these cavities.

The concentrations of the base cations extracted from the biochar samples are listed in Table S3 and S4 in ESI file while Table 2 reports the CEC-NH₄ obtained by the quantification of ammonium from ion chromatography (A) and colorimetric method (B). Both readily soluble cations and exchangeable cations after washing with ammonium acetate are considered. It has to be underlined that the concentrations of Fe³⁺, Al³⁺, and Mn²⁺ ions are very low in all the analyzed biochar (Table S4) and thus they are not considered to calculate the CEC-BC values.

The observed CEC-NH₄ values (as summarized in Fig. 8) indicate that poplar and willow biochars obtained at 400 °C exchange more cations than the corresponding product obtained from pine wood. The lowest CEC-NH₄ value is obtained for black pine char produced at processing temperature above 400 °C. Considering that this sample possesses the highest surface area, this result suggests that the surface area is not the main factor involved in the exchange of ions with the environment. Furthermore, the highest CEC values are observed for chars produced at pyrolysis temperature of 400 °C, which are the samples with the lowest surface area if compared to the samples produced at 550 °C and 650 °C.

The observed behavior can be therefore explained considering also the effect of high temperature on the polar functional groups on chars' surface. In fact, the increase of temperature leads to a decrease of polar components as evidenced by the percentage of oxygen + nitrogen reported in Fig. 8. This evidence indicates that CEC is mainly controlled by the rate at which the polar groups are removed rather than the new surface area formed increasing the pyrolysis temperature. Among the exchangeable cations, Ca²⁺ shows the highest tendency to be exchanged, especially from poplar and willow biochar obtained at a temperature of 550 °C.

In conclusion, we can summarize that the CEC values depend on two different parameters, temperature of pyrolysis, and starting feedstock.

- Pyrolysis temperature: the increase of the temperature leads to an increase of the surface area and a gradual removal of polar species (O and N). Probably, the decrease of polar compounds in the sample might be connected to the decrease of the CEC value.
- Starting feedstock: considering the biochars produced at the same temperature, the CEC values depend on the starting feedstock. In particular, we observed that black pine biochars possess the lowest CEC values even if they are characterized by higher surface area and higher

Q2 t2.1 **Table 2** CEC-NH₄ obtained from
t2.2 ion chromatography (A) and
t2.3 colorimetric method (B). The
t2.4 values in parentheses are the
t2.5 standard deviation on the last
significant figure

[cmol(+)/kg]	BP 400 °C	P	W	BP 550 °C	P	W	BP 650 °C	P	W
CEC-NH ₄ (A)	44(2)	48(3)	67(4)	16.7(8)	19.6(9)	19.8(9)	12.1(6)	15.9(8)	16.8(9)
CEC-NH ₄ (B)	45(4)	52(3)	65(5)	20(2)	20(1)	23(2)	12(1)	17(1)	18(1)

457 amounts of polar components. This behavior can be
458 explained considering that black pine chars exhibit
459 the presence of pores in the nanometric size range
460 that are not relevant for water exchange due to the
461 high capillary forces.

462 **4 Conclusions**

463 The aim of this study is to investigate at lab-scale the relation-
464 ship between the physical-chemical properties of the obtained
465 biochars and the production conditions, such as the biomass
466 feedstock and the pyrolytic temperature. Nine biochars were
467 pyrolyzed at 400, 550, and 650 °C from three different bio-
468 masses (black pine, poplar, and willow), selected from both
469 hard and softwood. Here the combined use of nitrogen adsorp-
470 tion isotherm, Hg porosimetry, electron microscopy, and eval-
471 uation of the cation exchange capacity (CEC) is proposed to
472 correlate the properties of the char with the production condi-
473 tion and starting biomass.

474 The results indicate that biochars with high surface area can
475 be obtained at high T, and this effect is more evident for pine
476 biochar, which shows the largest surface area (504 m²/g) at
477 550 °C. The pore analysis evidences that chars are character-
478 ized by two different types of pores: micro-pores in the range
479 between 1 and 10 nm that are not remarkably affected by the
480 starting feedstocks, together with macro-pores whose size is
481 strongly dependent on the morphological structure of the ini-
482 tial wood type. Indeed, poplar and willow samples pyrolyzed
483 at 550 °C possess pores in the range between 0.5 and 500 μm,
484 with a small population centered at around 1 μm, while pore
485 distribution for pine char shows a maximum at around 10 μm
486 together with larger pores with a diameter around 200 μm.

487 Regarding the retention/release properties of the investigat-
488 ed samples, our investigation demonstrates that poplar and
489 willow biochars exchange more cations than the correspond-
490 ing product obtained from black pine wood. This behavior can
491 be explained considering that black pine chars exhibit the
492 presence of pores in the nanometric size range that are not
493 relevant for water exchange due to the high capillary forces.
494 Furthermore, we also note that the highest CEC values are
495 observed for chars produced at pyrolysis temperature of
496 400 °C, which are the samples with the lowest surface area
497 if compared to the samples produced at 550 °C and 650 °C.

This finding suggests an effect of the temperature on the
final properties of these materials: indeed, the increase of
the pyrolysis temperature leads to a gradual removal of po-
lar species and, consequently, a decrease of the CEC is ob-
served for the samples produced at 550 °C and 650 °C. To
summarize, the cation exchange capacity, which is a very
important parameter when char is employed for soil appli-
cations, seems to be mainly dependent on the amount of
polar component on their surface. The effect of chars' sur-
face area has to be considered as well.

This study aims at representing a step forward in the
characterization of the char produced by pyrolysis of
biomass. Although previous reports investigated bio-
chars produced from different biomasses at different py-
rolysis temperatures, they only deal with the water re-
tention [20] or the mechanical properties [21] of the
obtained chars without focusing on the CEC. On the
other hand, works reporting the measure of the CEC
for different chars did not provide an exhaustive char-
acterization of the carbonaceous material [15, 18, 19] as
in the present study.

Here, we propose a multi-technique approach for a complete
characterization of the carbonaceous substrates in terms of chem-
ical composition, morphology, and porosity as a function of the
starting wood type and pyrolysis temperature. In addition, we
report on an optimized methodology for the evaluation of the
CEC in order to give an insight on the structure-property correla-
tion of the biochars. This experimental approach can be used to
gain additional information on the CEC capacity of the chars
helping in the optimization of the parameters used in the prepara-
tion of these materials. However, this work is based on TGA lab-
scale pyrolysis: further analysis with slow pyrolysis carried out at
pilot or demo scale will be needed, as biochars obtained in pilot
reactors are expected to be different from those obtained in TGA.
Moreover, other parameters will heavily impact on the product
characteristics, namely reactor design and solid residence time.

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Q3 541 **Author Contributions** David Chiamonti, Emiliano Fratini, Luca Rosi: 542 conceptualization. 543 Andrea Maria Rizzo, Lorenzo Bettucci, David Casini, Giovanni 544 Ferraro, Emiliano Fratini, Giuditta Pecori: formal analysis. 545 Giovanni Ferraro, Emiliano Fratini, Giuditta Pecori, Luca Rosi: data 546 curation. 547 Giovanni Ferraro, Emiliano Fratini, Luca Rosi: writing—original draft 548 preparation. 549 Lorenzo Bettucci, Giovanni Ferraro, Emiliano Fratini, Giuditta Pecori, 550 Luca Rosi: methodology. 551 David Chiamonti, Giovanni Ferraro, Emiliano Fratini, Luca Rosi: 552 writing—review and editing.

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559 Compliance with ethical standards

560 **Competing interests** The authors declare that they have no competing 561 interests. 562

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